

Original Research

Adsorption of Rhodamine B from Aqueous Solution by Goat Manure Biochar: Kinetics, Isotherms, and Thermodynamic Studies

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Abstract

Biochar was prepared by pyrolyzing goat manure at 600°C (GMB600) and used as an adsorbent for the removal of Rhodamine B (RhB) from aqueous solution. The structure and properties of GMB600 were characterized by elemental analysis, BET, SEM and FTIR. The effects of solution pH, biochar dosage, contact time, initial RhB concentration and temperature on the adsorption behaviors were investigated by batch experiments, which were further used to preliminarily discuss the adsorption mechanism. Results showed that the adsorption equilibrium could attain in 150 min and the RhB removal rate could reach 96.5% under the conditions of solution pH 5.0, GMB600 dosage of 0.4 g/L, initial RhB concentration of 20 mg/L and temperature of 25°C. The adsorption behavior of RhB onto GMB600 followed the pseudo second-order kinetic model, and the adsorption isotherm was better fitted with the Freundlich model. Thermodynamic parameters such as Gibbs free energy change (ΔG°), enthalpy change (ΔH°) and entropy change (ΔS°) indicated that the adsorption of RhB onto GMB600 was a spontaneous and endothermic process with increased entropy. FTIR analysis showed that the oxygen-containing functional groups on GMB600 provided RhB with abundant adsorption sites, and the adsorption mechanism was possibly attributable to the hydrogen bond and π - π interactions between RhB and GMB600.

Keywords: biochar, goat manure, Rhodamine B, adsorption kinetics, isotherms, thermodynamics

Introduction

Rhodamine B (RhB) is a synthetic organic dye widely used in textile, paper-making, leather manufacture, stained glasswork, the cosmetics

industry and many other industries due to its high tinting strength, bright color and high stability [1, 2]. These industries produce a great amount of RhB dye wastewater, with the characteristics of high colority, poor biodegradability and heavy toxicity [3]. The traditional biological method hardly meets the emission requirements, while directly draining into water would cause serious threats to the environment and human health. Previous experimental studies have

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proven that RhB can lead to carcinogenic, teratogenic and mutagenic effects when the hazardous substance irritates eyes or skin, which causes severe harm to the health of human beings and animals [4, 5]. Considering this, it is especially important to explore efficient and economical ways for RhB wastewater treatment.

At present, researchers have tried a number of methods to remove RhB dye from aqueous solution, such as adsorption [6, 7], the electro-Fenton process [8], photodegradation [9], sonochemical degradation [10] and membrane separation [11]. Among them, adsorption has been regarded as one of the most potential methods and widely favored by researchers because of its advantages such as convenient operating conditions, good treatment effects and an abundant source of raw materials. A variety of materials including bentonite [12], montmorillonite [13], activated carbon [14], as well as many other materials prepared by agricultural bio-wastes [15, 16] have been used as adsorbents for RhB removal from aqueous solutions. Moreover, it has become a research hotspot in recent years for exploring new adsorbents featuring high-efficiency, low-cost and non-secondary pollution.

Biochar is a type of carbon-containing product produced by biomass residual body through high-temperature pyrolysis in an oxygen-limited or oxygen-free state [17]. Compared with commercial activated carbon, biochar has a large specific area and good adsorption capacity as well. Moreover, it is much cheaper than activated carbon owing to its simple preparation process [18]. In addition, biochar can also effectively improve the physical and chemical properties of soil and further increase crop production [19]. Therefore, reports concerning the use of bio-char to absorb organic dyes have been increasing recently. In previous studies, straw [20, 21], bark [22], water hyacinth [23] and some other plant source materials were used to prepare biochar. China is one of the largest farming countries with an enormous amount of animal manure produced by livestock and poultry farming, which would cause great pressure on the environment if these wastes could not be effectively treated or recycled. Nevertheless, studies on the preparation of biochar from animal manure source materials for RhB dye adsorption have rarely been reported. The adsorption mechanism of RhB dye onto animal manure biochar is not yetThe objective of this study was to investigate the feasibility of biochar prepared from goat manure as a novel adsorbent for the removal of RhB dye from aqueous solutions. The effects of solution pH, biochar dosage, contact time, initial RhB concentration and temperature on the adsorption performance were studied. Based on this, the mechanism of RhB adsorption by goat manure biochar was further analyzed via kinetic, isothermal and thermodynamic models. The results would provide a theoretical basis for the application of goat manure biochar in RhB wastewater treatment and explore a new way for the resource utilization of goat manure.

Experimental

RhB Dye Solution

RhB, molecular formula $C_{28}H_{31}ClN_2O_3$, also known as basic rose red or brilliant pink B, was purchased from Aladdin Chemical Reagent Co. Ltd. The structural formula of RhB is shown in Fig. 1. The stock solution of RhB (1000 mg/L) was prepared in deionized water, while fresh dilutions of the desired RhB concentrations were prepared according to the experimental requirement. The pH value of RhB solution was adjusted by adding 0.1 mol/L HCl or 0.1 mol/L NaOH. The chemical reagents used in the experiment were all analytically pure.

Preparation of Goat Manure Biochar

Goat manure was collected from a farm in Ya'an city, Sichuan Province of China. The goat manure was air-dried under ambient conditions and ground to pass a 80 mesh sieve. The ground goat manure was put into a crucible, compacted and covered well. Then it was heated to 600°C in a muffle furnace at a 15°C per minute gradient and kept at constant temperature for 150 min. After the pyrolysis products cooled to room temperature, 20 mL of HCl (1.0 mol/L) was added to each and every 1 gram of the pyrolysis products and the solution was shaken at 150 rpm for 30 min. The operation was repeated 3 times to remove ash from the pyrolysis product. After filtration, the product was washed to neutral with deionized water, oven-dried at 105°C for 48 h and then passed through a 100 mesh sieve. The goat manure biochar (GMB600) was successfully prepared and then packed in a sealing bag for later use.

Adsorption Experiments

The effect of initial solution pH on RhB removal was studied over the pH range of 3.0–9.0 with 0.5 g/L of GMB600 for 300 min in a RhB solution of 20 mg/L concentration at 25°C. The effect of biochar dosage on RhB removal was investigated with different GMB600 dosage (0.2–2.0 g/L) in a RhB solution of 20 mg/L, pH 5.0 and shaken for 300 min at 25°C. The effect of contact time on RhB removal was investigated with GMB600 in a RhB solution of 20 mg/L, pH 5.0 and shaken for

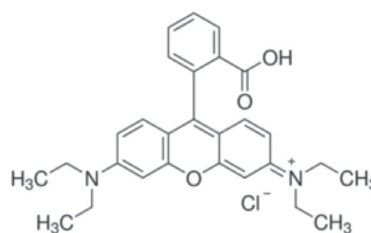


Fig. 1. Structural formula of RhB.

different adsorption time (5~500 min) at 25°C. The effect of initial RhB concentration and temperature on RhB removal was investigated by shaking 0.4 g/L of GMB600 for 150 min in RhB solution of different initial concentrations (10~100 mg/L), pH 5.0 at 298 K, 308K and 318 K, respectively.

All the above experiments were conducted in a gas bath homoiothermy shaker at 200 rpm. After adsorption, the suspensions were centrifuged at 4000 rpm for 5 minutes and then the prepared samples for analyses (4 mL) were filtered in a 0.45 µm membrane.

Analysis Methods

The contents of C, H, N and O of GMB600 were tested by elemental analyzer (VARIO EL cube, Elementar, German). The specific surface area, total pore volume and average pore diameter of GMB600 were tested using BET analysis (NOVA4000e, Quantachrome, America). The surface morphology of GMB600 was analyzed by a field emission scanning electron microscope (SUPRA40, ZEISS, German). The functional group on the surface of GMB600 was determined by a Fourier infrared spectrometer (FTIR Spectrum100, Perkin Elmer, America). The concentration of RhB was measured at λ_{\max} 554 nm by UV-Vis spectrophotometry (UV-1800PC, HITACHI, Japan). The mean data value obtained after three parallel tests was analyzed and fitted by origin 9.0.

The RhB removal rate (R , %) and amount of RhB adsorbed by GMB600 (q_e , mg/g) at equilibrium were calculated by the following equations:

$$R = \frac{C_0 - C_e}{C_0} \times 100\% \quad (1)$$

$$q_e = \frac{(C_0 - C_e)V}{m} \quad (2)$$

...where C_0 and C_e are the initial and equilibrium concentrations of RhB solution (mg/L), V is the volume of RhB solution (mL) and m is the mass of GMB600 (g).

Results and Discussion

Biochar Properties

The elementary composition of GMB600 is shown in Table 1. As seen, GMB600 contained a lot of C and a little bit of H and N. In general, the aromaticity and polarity of adsorption materials could be judged according to the ratio of H/C and O/C. The aromaticity of biochar decreased with the increase of H/C ratio, and the polarity of biochar increased with the increase of O/C ratio. The previous studies showed that the biochar contained concentrated aromatic rings when the H/C ratio was less than 0.6 and the half-life cycle of biochar was estimated at more than 1000 years

Table1. Elemental composition and BET analysis of GMB600.

Property	GMB600
Ash content (%)	18.37
C (%)	63.25
H (%)	2.21
O (%)	12.09
N (%)	2.49
H/C	0.035
O/C	0.191
pH	7.0
Specific surface area (m ² /g)	181.76
Total pore volume (cm ³ /g)	0.245
Average pore diameter (nm)	11.91

when the O/C ratio was less than 2.0 [24, 25]. In this study, the ratio of H/C and O/C was 0.035 and 0.191 respectively, indicating that GMB600 was of relatively high aromaticity and stability.

The SEM of GMB600 is given in Fig. 2. The surface of GMB600 was rough and uneven and its interior was loose and porous, with large specific surface area and rich pore structure. Additionally, according to the results of BET analysis shown in Table 1, the specific surface area and total pore volume of GMB600 were respectively 181.76 m²/g and 0.245 cm³/g, which would provide favorable conditions for RhB adsorption onto the biochar.

Effect of Operational Variables on RhB Adsorption

Effect of Initial Solution pH

pH value is an important factor affecting adsorption performance, since it can alter not only the charge or

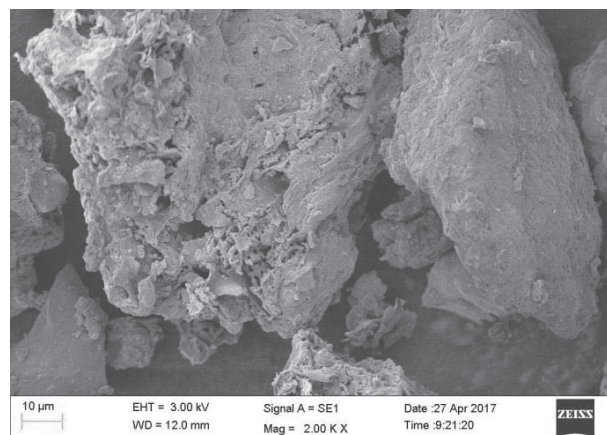


Fig. 2. Scanning electron microscope of GMB600.

species of adsorbate but also the surface properties of the adsorbent. The effect of initial solution pH on RhB adsorption by GMB600 is shown in Fig. 3a). As seen, the removal rate of RhB was over 97% in the pH of 4.0 to 5.0. Under acidic conditions, RhB mainly existed in the form of quaternary ammonium cation and molecule and the surface functional groups of biochar protonated [26]. So it was beneficial to obtain better adsorption rates via ion-exchange within the above pH range. When the pH value was 3.0, the surplus H^+ in the solution began to compete with RhB for the surface adsorption sites of biochar, resulting in a decrease in the amount of RhB adsorbed.

When the pH value was higher than 5.0, the removal rate of RhB onto GMB600 decreased with the increase of pH, and it was only 67.29% when the pH value increased to 10.0. Since the occurrence of carboxyl ionization in RhB in alkaline condition, the quaternary ammonium cation and carboxyl anions coexisted in the solution, causing polymerization and precipitation and leading to the blockade of the channel where RhB entering into the micro-pores of biochar [27, 28]. In addition, the excessive OH^- in the solution was able to compete for the surface adsorption sites of biochar with RhB, leading to a significant decrease in RhB removal rate.

Effect of Biochar Dosage

The effect of biochar dosage on RhB adsorption is shown in Fig. 3b). As illustrated in Fig. 3b), the removal rate of RhB increased with the increase of GMB600 dosage and then approached to a flat because

the increase of adsorbent dosage could provide a larger specific surface and more adsorption sites for RhB. In contrast, the amount of RhB adsorbed by unit mass of GMB600 decreased with the increase of biochar dosage. A decrease of q_e from 77.059 to 9.979 mg/g was observed when the GMB600 dosage increased from 0.2 to 2.0 g/L. The reason could be as follows. In the case of limited adsorbate, increasing biochar dosage could lead to a surplus of biochar, in turn resulting in a decrease of RhB adsorbed by unit mass of GMB600. This regulation was similar to the result of Aljeboree's research in crystal violet dye adsorption by activated carbon [29]. Taking both adsorption efficiency and economy into consideration, the optimal GMB600 dosage for the subsequent experiments was determined to be 0.4 g/L. Under this dosage condition, the removal rate of RhB was 96.89% and the q_e was 48.443 mg/g.

Effect of Contact Time

Fig. 3c) shows the effect of contact time on adsorption performance of GMB600. As can be seen from Fig. 3c), the adsorption process of RhB onto GMB600 could be divided into three phases, namely the rapid adsorption phase (phase I: 0~60 min), the slow diffusion phase (phase II: 60~120 min) and the adsorption equilibrium phase (phase III: 120~500 min).

In phase I, the adsorption sites of GMB600 were sufficient, RhB could rapidly occupy these sites under the action of mass transfer driving force and thus it could be adsorbed by GMB600 quickly. The RhB removal rate reached 86.25% in this phase. In phase II, the adsorption sites of GMB600 and the residual

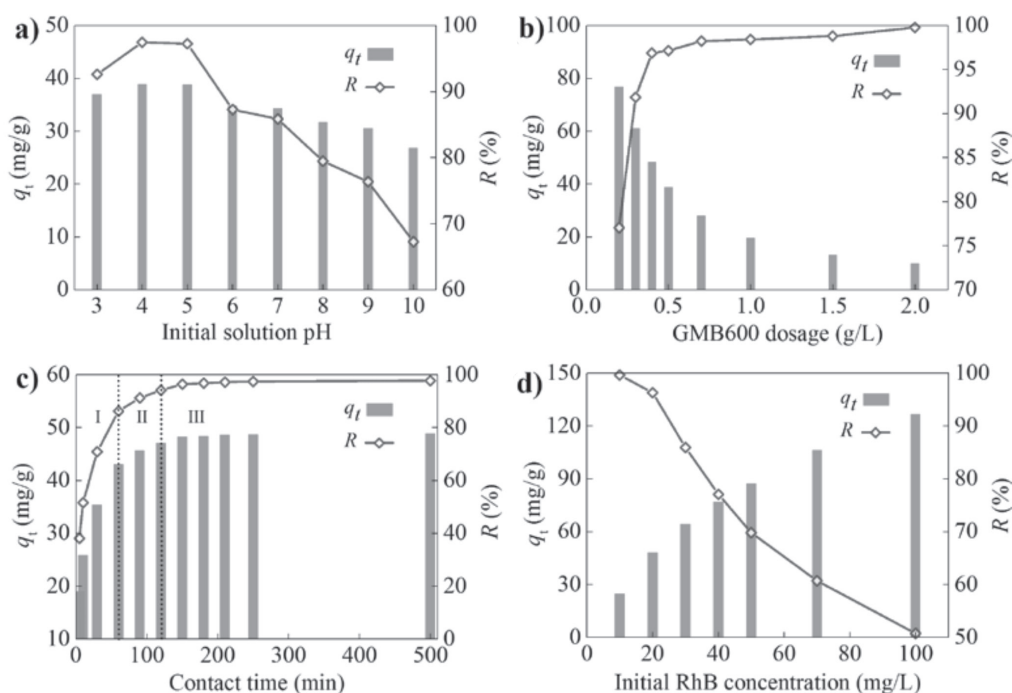


Fig. 3. Effect of initial solution pH a), GMB600 dosage b), contact time c) and initial RhB concentration d) on adsorption performance.

RhB in the solution were gradually reduced as the extension of contact time, resulting in a decrease in RhB concentration gradient between the GMB600 and solution. Thus, the adsorption rate of RhB went down with the reduction of mass transfer driving force. In phase III, the adsorption amount of RhB was almost unchanged as time prolonged. The adsorption of GMB600 for RhB reached equilibrium at 150 min, and the RhB removal rate reached 96.5%. Therefore, the optimal contact time for the subsequent experiments was determined to be 150 min.

Effect of Initial RhB Concentration

The adsorption efficiency of GMB600 under different initial RhB concentrations is depicted in Fig. 3d). The results showed that, with the increase of the RhB initial concentration from 10 mg/L to 100 mg/L, the equilibrium adsorption amount of RhB increased from 24.652 mg/g to 126.828 mg/g, while the corresponding removal rate showed a decrease of from 99.65% to 50.73%. Analysis suggested that when the initial RhB concentration increased, the number of RhB molecules distributed around the biochar increased correspondingly. As such, the RhB concentration gradient between the GMB600 and solution increased, leading to an increase of mass-transfer force so that more RhB could be adsorbed by GMB600. However, since the GMB600 dosage was constant, the adsorption sites for RhB was also limited. With the increase of the initial RhB concentration, the remaining unadsorbed RhB in the solution also increased, resulting in the decrease of RhB removal rate.

Adsorption Kinetics

To further study the adsorption process of RhB onto GMB600, the adsorption data were analyzed using four well-known kinetic models, specifically the pseudo first-order model [30] (Eq. 3), pseudo second-order model [31] (Eq. 4), intra-particle diffusion model [32] (Eq. 5) and Elovich model [33] (Eq. 6):

$$\ln(q_e - q_t) = \ln q_e - k_1 t \quad (3)$$

$$\frac{t}{q_t} = \frac{t}{q_e} + \frac{1}{k_2 q_e^2} \quad (4)$$

$$q_t = k_3 t^{0.5} + C \quad (5)$$

$$q_t = \left(\frac{1}{\beta}\right) \ln(\alpha \cdot \beta) + \left(\frac{1}{\beta}\right) \ln(t) \quad (6)$$

...where q_e is the equilibrium adsorption amount of RhB (mg/g), q_t is the adsorption amount of RhB at time t (mg/g), k_1 is the pseudo first-order rate constant (min^{-1}), k_2 is the pseudo second-order rate constant ($\text{g}/(\text{mg} \cdot \text{min})$), k_3 is the intra-particle diffusion rate constant ($\text{mg}/(\text{g} \cdot \text{min}^{0.5})$), C is the intercept, α is the initial adsorption rate constant ($\text{g}/(\text{mg} \cdot \text{min})$), β is the desorption rate constant (g/mg), and t is the contact time (min).

The adsorption kinetics curves of RhB (20 and 50 mg/L) onto GMB600 are given in Fig. 4. The corresponding fitted parameters are listed in Table 2. We can see that the coefficient of determination of the

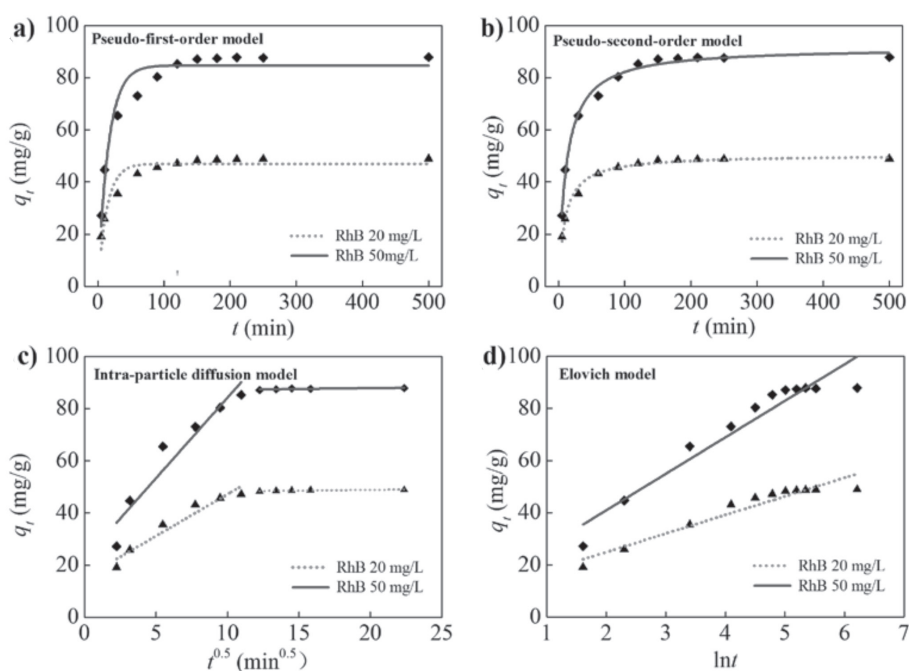


Fig. 4. Adsorption kinetics curves of RhB onto GMB600: a) pseudo first-order model, b) pseudo second-order model, c) intra-particle diffusion model, d) Elovich model.

Table 2. Fitting parameters of kinetic models for RhB adsorption onto GMB600.

Kinetic model	Parameter	Initial RhB concentration (mg/L)	
		20	50
Pseudo-first-order	q_e (mg/g)	47.026	84.628
	k_1 (min ⁻¹)	0.0725	0.0630
	R^2	0.9092	0.9369
Pseudo-second-order	q_e (mg/g)	50.436	91.645
	k_2 (g/(mg·min))	0.0021	0.0001
	R^2	0.9871	0.9913
Intra-particle diffusion	C_1	15.216	22.460
	k_{3-1} (mg/(g·min ^{0.5}))	3.1960	6.1734
	R^2	0.9271	0.8985
	C_2	47.745	86.645
	k_{3-2} (mg/(g·min ^{0.5}))	0.0538	0.0580
	R^2	0.6674	0.4571
Elovich	α (g/(mg·min))	31.662	35.286
	β (g/mg)	0.1402	0.0715
	R^2	0.9198	0.9158

pseudo second-order model ($R^2 > 0.98$) was obviously greater than that of other models. The theoretical q_e values obtained from the pseudo second-order model were very close to the experimental q_e values, and their relative error was only 3.19% and 4.32% respectively. Thus, the pseudo second-order model could fit the adsorption data better than other models.

The pseudo second-order model reflected the whole adsorption process of RhB onto GMB600, such as liquid film diffusion, surface adsorption and intra-particle diffusion [34]. Additionally, according to the fitting results of the intra-particle diffusion model, the whole adsorption process is divided into two stages. The first stage is the diffusion of RhB across the liquid membrane to the surface of GMB600 (membrane diffusion), and the second stage is the diffusion of RhB in pores of GMB600 (internal diffusion). The facts that $k_{3-1} > k_{3-2}$ and $C_1 < C_2$ showed that the diffusion of RhB on the surface of GMB600 was faster, and the main speed-limiting step in the adsorption process was intra-particle diffusion [3]. The Elovich model could describe the diffusion of adsorbents at the adsorption interface, surface activation and deactivation. It is suitable for those reaction processes in which the activation energy varies greatly and can reveal the data irregularity which is ignored by other types of kinetic models. In this research, the coefficient of determination of the Elovich model was less than 0.92, which is much less than that of the pseudo second-order model. This indicated that the Elovich model was unfit for the adsorbent process in this study.

Adsorption Isotherms

Adsorption isotherms have been widely applied to evaluate the maximum capacity of adsorbents and to reveal the essence of adsorbate-adsorbent interactions. In the present study, the adsorption data were fitted by the Langmuir [35] and Freundlich [36] models, which can be expressed as Eqs. (7) and (8), respectively:

$$q_e = \frac{q_m K_L C_e}{1 + K_L C_e} \quad (7)$$

$$q_e = K_F C_e^{1/n} \quad (8)$$

...where q_e and q_m (mg/g) are the equilibrium and theoretical maximum adsorption capacities, respectively, C_e is the RhB concentration at equilibrium (mg/L), K_L is the adsorption equilibrium constant of the Langmuir model (L/mg), K_F is the adsorption capacity parameter of the Freundlich model, and $1/n$ is the index of the Freundlich model.

The adsorption isotherms of RhB onto GMB600 are displayed in Fig. 5 and the relevant fitting parameters are presented in Table 3. It could be seen that the Freundlich model could fit the adsorption data better than the Langmuir model with the R^2 values above 0.986, indicating that the adsorption of RhB onto GMB600 was multilayer adsorption. Generally, the constant K_F was positively related to the adsorption capacity. In this study, K_F increased with the increase of temperature, indicating that the rise of temperature

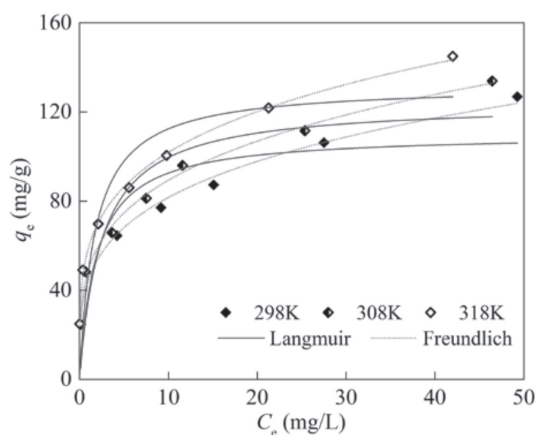


Fig. 5. Isotherm fitting curves of RhB adsorption onto GMB600.

could improve adsorption capacity. The constant $1/n$ was positively related to surface homogeneity. The adsorption was difficult to happen if the $1/n$ was greater than 1.0, while it was relatively easy to happen if the $1/n$ was less than 0.5 [37]. In this study, the values of $1/n$ were all lower than 0.3, indicating that the RhB could be easily adsorbed by GMB600. In addition, the value of $1/n$ decreased with the increase of temperature, indicating that the rise of temperature could make the adsorption happen more easily.

Table 4 showed the K_F values of different adsorbents for the RhB adsorption process. As seen, there are significant differences in the K_F values when different adsorbents adsorbed the same dye. The K_F values in this study were much higher than for other adsorbents, indicating that GMB600 had stronger adsorption capacity for RhB than other adsorbents. GMB600 has great potential application in the elimination of RhB.

Adsorption Thermodynamics

Thermodynamic parameters such as Gibbs free energy change (ΔG°), enthalpy change (ΔH°) and entropy change (ΔS°) for RhB adsorption onto GMB600 were calculated as Eqs. (9) and (10):

$$\Delta G^\circ = -RT \ln K_d \quad (9)$$

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ \quad (10)$$

...where R is the ideal gas constant (8.314 J/(mol·K)), T is the Kelvin temperature (K), ΔG° is the Gibbs free energy change (kJ/mol), ΔH° is the enthalpy change (kJ/mol), ΔS° is the entropy change (kJ/(mol·K)), and K_d is the thermodynamic equilibrium constant.

The calculation results are illustrated in Table 5. The negative values of ΔG° implied that the adsorption process of RhB onto GMB600 was spontaneous [43]. The decrease of ΔG° as temperature increased indicated that high temperature could promote the spontaneous trend of adsorption, which is consistent with the previous results obtained from the Freundlich model. The positive value of ΔH° (22.5519 kJ/mol) revealed that the adsorption would be an endothermic process [44]. According to previous studies [45, 46], the adsorption process was generally considered as chemical adsorption if $\Delta H^\circ > 40$ kJ/mol or physical adsorption if $\Delta H^\circ < 25$ kJ/mol. Thus, the RhB adsorption onto GMB600 was primarily physical process. As reported by Tan et al. [47], the dominant force of adsorption was van der Waals forces if ΔH° ranged from 4 to 10 kJ/mol, hydrogen bonding force if ΔH° ranged from 2 to 40 kJ/mol and dipole bonding force if ΔH° ranged from 2 to 29 kJ/mol. Therefore, it could be speculated

Table 3. Fitting parameters of isotherm models for RhB adsorption onto GMB600.

T (K)	Langmuir model			Freundlich model		
	q_m	K_L	R^2	K_F	$1/n$	R^2
298	110.078	0.5284	0.7524	44.697	0.2615	0.9864
308	124.072	0.4071	0.7920	49.272	0.2585	0.9919
318	132.034	0.5723	0.7801	58.872	0.2382	0.9979

Table 4. K_F values of different adsorbents for the RhB adsorption process.

Adsorbent	K_F	References
Beta zeolites	2.54	CHENG et al. [38]
<i>Pongamia glabra</i> seed cover biochar	4.653	BORDOLOI et al. [39]
Cotton stalks-activated carbon	6.096	ÖZDEMİR et al. [40]
Chicken bone-based biochar	2.98–6.23	OLADIPO et al. [41]
Earthworm manure derived biochar	2.53–8.70	WANG et al. [42]
GMB600	44.697–58.872	This study

Table 5. Thermodynamic parameters of RhB adsorption onto GMB600.

$T(K)$	K_d	ΔG°	ΔH°	ΔS°	R^2
298	5.783	-4.348	22.5519	0.0904	0.9787

that the adsorption of RhB onto GMB600 primarily involved a hydrogen and/or dipole bonding force. The positive value of ΔS° (0.0904 kJ/(mol·K)) indicated that the adsorption was a process of entropy increase. In this process, the order degree of the solid-liquid interface decreased [48].

FTIR Analysis Before and After Adsorption

To further analyze the adsorption mechanism, the infrared spectrum analysis was conducted to identify the functional groups on the surface of GMB600 before and after adsorption, and the results are presented in Fig. 6.

As observed from Fig. 6, the GMB600 before adsorption contained a number of oxygen-containing groups such as hydroxy, carbonyl and ester, which provided favorable conditions for the adsorption of RhB. The adsorption peaks at 3673, 1574, 1047 and 794 cm^{-1} were attributed to the stretching vibration of hydroxyl groups (-OH), the bending vibration of C=C and C=O in aromatic, the stretching vibration of C-O in esters and ethers, and the bending vibration of C-H in aromatic and heterocyclic [49], respectively.

After the adsorption of RhB, several new peaks appeared at 3020–3384 and 1401 cm^{-1} , which were the adsorption peaks of -OH and C-H, respectively. The adsorption peak of C-O shifted to 1062 cm^{-1} , indicating that C-O was involved in the adsorption process of RhB onto GMB600. The adsorption peak of -OH shifted to 3686 cm^{-1} , suggesting that the hydrogen bond might be formed via the combination of -CH₃, -COOH in

RhB and the oxygen-containing functional groups on GMB600. Therefore, the hydrogen bond played an important role in the adsorption process, which was in agreement with the result of the above thermodynamic analysis [50]. Meanwhile, the adsorption peak at 1574 cm^{-1} shifted to 1577 cm^{-1} and the adsorption peak at 794 cm^{-1} disappeared. This may be associated with the π - π bond formed via the combination of π electron acceptor (-COOH) in RhB and π electron donors (C=C, C-H) on GMB600 [51]. Thus, the π - π bond was also of great significance to the adsorption of RhB onto GMB600.

Conclusions

In the present study, the biochar prepared by goat manure at 600°C was found to be a potential adsorbent for the removal of RhB from aqueous solution. The RhB was optimally adsorbed at pH 5.0, GMB600 dosage 0.4 g/L and contact time for 150min when initial RhB concentration was 20 mg/L and temperature was 25°C. The adsorption process of RhB onto GMB600 was found to obey the pseudo second-order model and the Freundlich model well. The fitting results showed that the intra-particle diffusion was not the only limiting step, liquid film diffusion and surface adsorption may also regulate the adsorption rate. The thermodynamic study showed that the adsorption of RhB onto GMB600 was a spontaneous, endothermic and entropy-increasing process. According to the FTIR analysis, the adsorption of RhB onto GMB600 was closely related to the effect of hydrogen bond and π - π bond.

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Conflict of Interest

The authors declare no conflict of interest.

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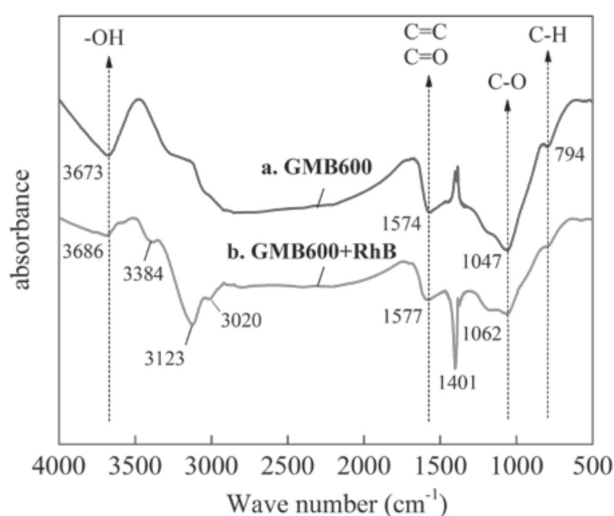


Fig. 6. FTIR spectra before a. and after b. RhB adsorption by GMB600.

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