Introduction

Factors such as increasing population, urbanization and industrialization in the last century threaten marine ecosystems, especially coastal areas. Many studies have been carried out using different methods to determine the effects of industrial and anthropogenic waste on marine ecosystems [1-6]. As waste oil is treated in the harbors that contain various waste separation activities such as Kepez Harbor in the Dardanelles, it is expected that there will be oil-grease input in the system if it is not treated properly. Such a situation may cause pollution in the receiving environment, disrupting the structure of both the benthic and the pelagic ecosystems. In developed countries, various government agencies...

Potential Impacts of a Waste Reception Facility on Stable Isotope Composition of Zooplankton and Particulate Organic Matter in the Dardanelles

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Abstract

In this study, seasonal stable isotope compositions of zooplankton and particulate organic matter were analyzed from the coastal zone in Kepez Harbor, the discharge point of a waste treatment plant. The results were compared with the results of a control station located 10 km away from the harbor. Zooplankton samples were collected horizontally with a 200 µm mesh plankton net from the surface waters. POM samples were collected with a 5L Nansen bottle and filtered through precombusted GF/F filters. Variables such as temperature, salinity, chlorophyll-a as well as oil-grease and chemical oxygen demand were also determined in the study sites. The salinity values in the control station were significantly higher than the harbor station through the whole year. The δ¹³C values of POM in the control station were significantly higher than the values observed in the harbor station. The δ¹⁵N values of zooplankton in the harbor were higher than the values obtained in the control station; however, the differences were not significant. The values of oil-grease obtained in two stations were lower than the values pronounced in EPA, EU, and Turkish Regulations. The study showed that the stable isotope compositions were affected by terrestrial inputs and high currents rather than the facilities comprised in the harbor.

Keywords: δ¹³C, δ¹⁵N, water pollution, food web, seasonal variation

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or non-governmental organizations determine the limit values that can be discharged into the receiving environment with various directives and regulations. Within the scope of the coastal ecosystem, keeping such situations under control is left to the responsibility of industrial facilities. In Turkey, the facilities that discharge to marine systems or inland waters also control the effects of the treatment plants to the aquatic systems by making routine analyses of the variables such as heavy metals, oil-grease, biological oxygen demand (BOD) and chemical oxygen demand (COD) in their facility laboratories or other accredited ones. However, there may be some amount of inorganic waste input into the receiving system even though it may be below the limit values provided by the international or the national organizations. Thus, stable isotope analysis can be an effective method to determine the possible effects of such conditions on marine systems.

Stable isotopes have been used extensively in order to determine the effects of waste and terrestrial inputs discharged into aquatic environments, sources of pollution and food web interactions [7, 8]. However, the number of studies in the coastal areas and in the regions where industrial activities may affect the quality of the coastal ecosystem is very low [9]. Mesozooplankton play an important role in transferring organic matter from phytoplankton and other organic materials (POM) to the higher trophic levels due to their trophic position. For this reason, they also have a role in the transport of pollution to the upper layers in the marine system. In marine ecosystems, δ13C and δ15N isotopes are frequently used to determine the sources of organic matter and their transfer through the trophic level [10, 11]. Consequently, consumer’s tissue δ15N can be used to estimate the trophic levels of organisms because they typically show 2.5-4.5% enrichment relative to their prey [12]. Conversely, the δ13C ratio in the tissue of the organism is enriched very weakly through higher trophic levels with 0.4-2.0‰, which can act as a useful indicator of the primary organic carbon source of an organism’s food intake [13]. Therefore, when δ13C is used to determine the nutritional carbon source of organisms, δ15N is used to estimate the trophic level that they may belong to.

The present study was conducted in Kepez Harbor (Port of Çanakkale) in the Dardanelles, where Çanakkale Strait plays a crucial role in transporting nutrients, plankton and fish communities between the Black Sea and the Mediterranean [14]. It is known that there are daily changes in nutrients as a result of discharges and terrestrial inputs in the Dardanelles [15-17]. In addition, the strait has a special two-layer flow regime in which the upper layer of the water column is composed of waters with low salinity level from the Black Sea, while the lower layer consists of waters with high salinity originating from the Mediterranean [18]. The aim of this study was to investigate the effects of the water discharged from the waste reception facility in Kepez Harbor to the marine system of the area. The objectives were: (i) to determine and compare the zooplankton and POM stable isotope compositions of the station in the harbor and the control station, (ii) to determine the differences of physico-chemical and biological factors between these stations (iii) and to determine the possible causes of these differences.

Material and Methods

The study was carried out in two stations in the Dardanelles between Autumn 2016 and Summer 2017. The first station is the coastal region where the waste water treatment plant belonging to the waste reception facility operating in Kepez Harbor has a discharge (Fig. 1). This station may occasionally be under the influence of strong currents, it is closer to the settlement and can be affected from terrestrial inputs. The depth of the station varies from 3 to 3.5m. The 2nd station is the control station and is a coastal area about 10 km southwest of the first station (Fig. 1). It has no terrestrial inputs and is less affected by the currents compared to the first station, and its depth varies between 2-2.5 m.

Temperature and salinity values were determined in situ using a YSI 650 MPS. Water samples for chlorophyll-a (Chl-a) measurement were collected from the surface with a 5L Nansen bottle and filtered through a 47mm diameter GF/F. Spectrophotometric determination was conducted according to [19] after 90% acetone extraction. Total suspended solids (TSS) determination was performed gravimetrically after the water samples were filtered through GF/C filters [20]. Chemical oxygen demand (COD) and oil-grease analysis were conducted using the open reflux method and standard methods for the examination of water and wastewater according to [21].

Samples for the stable isotope analyses (SIA) of POM were collected from the two stations with a 5 L Nansen bottle. The samples were first filtered through a 50µm plankton net to remove large organisms, then filtered again through precombusted (400°C, 3 hours) GF/F filters and dried in a drying oven at 60°C for 24 hours. Zooplankton samples for SIA were collected horizontally with a 200µm pore-sized plankton net from surface waters of the two stations. 200µm- to 1000µm-sized zooplankton were targeted, and larger organisms were removed from the samples under the microscope. Then bulk zooplankton samples were kept in -80°C until SIA. Before analysis the bulk samples were dried in a drying oven at 60°C for 24 hours and each sample was weighed (0.150mg) in tin capsules. The SIA for zooplankton and POM samples were conducted using an isotope ratio mass spectrometer (Delta V Thermo Finnigan; Thermo, Waltham, USA) in the Akdeniz University Food Safety and Agricultural Research Center. Since acidification can affect the nitrogen and carbon isotope values, the samples were not acidified.
The obtained results are expressed by $\delta$ using the following equation:

$$\delta^{13}\text{C or } \delta^{15}\text{N (‰)} = (R_{\text{sample}}/R_{\text{standard}} - 1) \times 1000$$

...where the stable isotope abundance of C and N was expressed as $\delta^{13}\text{C or } \delta^{15}\text{N}$, $^{13}\text{C}/^{12}\text{C}$ or $^{15}\text{N}/^{14}\text{N}$. The stable isotope ratio was represented by $R$ with the reference being V-PDB (Vienna Pee Dee Belemnite) and atmospheric $N_2$ isotope standards.

To determine the spatial differences between physico-chemical variables and the stable isotope ratios of zooplankton and POM, one-way analyses of variances (ANOVA; significance level $p < 0.05$) were used. The different groups were analyzed using Tukey’s honestly significant difference (HSD) test for unequal
sample sizes for post hoc comparisons among stations. Pearson’s correlation analyses were applied to find out whether there is a linear relationship between variables observed in the study area. For the analyses we used SPSS 18.0 and Minitab 17.1.0.

**Results and Discussion**

Results of the one-way ANOVA (significance level \( p<0.05 \)) determined during the study are shown in Table 1. The water temperature values in the study area followed a seasonal cycle, the lowest value being 7.41ºC in winter and the highest value being 25.45ºC in summer in the harbor station (Fig. 2). There was no significant difference in temperature between stations (Table 1; \( p>0.05 \)). Salinity values of the control station were significantly higher than the harbor station (Fig. 2, Table 1; \( p<0.001 \)). The highest salinity value was 30.38‰ and it was determined in the control station in summer, while the lowest value was 24.18‰ in the harbor station. Chl-\( a \) values were in the range of 1.26-2.90 μg L\(^{-1}\) in the harbor and 1.53-7.21 μg L\(^{-1}\) in the control station. Chl-\( a \) values were statistically significant among stations (Fig. 2, Table 1; \( p<0.05 \)). In terms of TSS, high values were observed in the control station compared to the harbor and the differences between the stations were statistically significant (Table 1; \( p<0.05 \)). While chl-\( a \) values showed significant positive correlation with TSS in the control station, there wasn’t a significant relation in the harbor (Table 2). The differences between stations were not found to be significant in COD values, while both the highest (in spring) and the lowest (in winter)
values were determined in the control station (Table 1; \( p > 0.05 \)). Oil and grease amounts were lower than 0.1 mg L\(^{-1} \) at all stations throughout the sampling period (Table 1).

Although COD values did not exceed the limit values as defined by the WCPR all through the sampling period, it was observed that the limit values of COD as defined by EU and EPA were exceeded during the spring and the summer samplings in the harbor station, and the autumn and the spring samplings in the control station (Table 3). The oil-grease values did not exceed the limit values defined in the EPA, EU and WCPR throughout the year (Table 3). Recently, no research has been conducted on COD in the marine system of the region, and the previous studies carried out were mostly in inland waters. For example, in a study conducted in 2001 [23], the COD values in Sarıçay River, which flows through the Strait, were found to be between 62 and 88 mg L\(^{-1} \), showing that this river plays an important role in transporting pollutants to the coastal system.

Stable carbon isotope values of zooplankton (hereafter \( \delta^{13}C_{zoo} \)) at the harbor station showed lower values in autumn and winter than in spring and summer, when surface water temperatures were high. Spatial variations of \( \delta^{13}C_{zoo} \) were not statistically significant (Table 1; \( p > 0.05 \)). The stable carbon isotope values of POM (hereafter \( \delta^{13}C_{POM} \)) determined at the control station were significantly higher than the harbor station (Table 1; \( p < 0.05 \)). Even the stable isotope nitrogen values of zooplankton \( \delta^{15}N \) (hereafter \( \delta^{15}N_{zoo} \)) showed higher variations in the harbor station than in the control, the differences between stations were not significant (Table 1; \( p > 0.05 \)). The variations of the stable nitrogen isotope values of POM (hereafter \( \delta^{15}N_{POM} \)) between stations were not significant (Table 1; \( p > 0.05 \)). TSS values showed significant positive correlation with \( \delta^{15}N_{zoo} \) and \( \delta^{15}N_{POM} \) in the harbor station (Table 2).

In another study conducted in the Mediterranean coastal system, it was found that anthropogenic and industrial wastes increased COD values in the marine system [24]. The results obtained in both studies showed similarity with the present study. On the other hand, COD values obtained in a study conducted in a wastewater treatment plant in the Marmara Region were found to be significantly higher than the COD values obtained in our study [25], while in another study conducted in the mouth of a river in the Black Sea the values were found to be very low [26].

Table 2. Pearson linear correlations between physico-chemical variables and stable isotope values of zooplankton and POM at the study site.

<table>
<thead>
<tr>
<th>Variables</th>
<th>Harbor</th>
<th>Control</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>( r )</td>
<td>( p )</td>
</tr>
<tr>
<td>chl-a vs TSS</td>
<td>0.542</td>
<td>n.s.</td>
</tr>
<tr>
<td>chl-a vs ( \delta^{13}C_{zoo} )</td>
<td>0.540</td>
<td>n.s.</td>
</tr>
<tr>
<td>chl-a vs ( \delta^{13}C_{POM} )</td>
<td>-0.529</td>
<td>n.s.</td>
</tr>
<tr>
<td>( \delta^{13}C_{zoo} ) vs ( \delta^{13}C_{POM} )</td>
<td>-0.305</td>
<td>n.s.</td>
</tr>
<tr>
<td>TSS vs ( \delta^{15}N_{zoo} )</td>
<td>0.782</td>
<td>&lt;0.05</td>
</tr>
<tr>
<td>TSS vs ( \delta^{15}N_{POM} )</td>
<td>0.666</td>
<td>&lt;0.05</td>
</tr>
</tbody>
</table>

n.s.: Not significant

Table 3. Limit values for water quality variables in national and international organizations.

<table>
<thead>
<tr>
<th>Variable</th>
<th>WCPR(^{1} )</th>
<th>EU(^{2} )</th>
<th>EPA(^{3} )</th>
</tr>
</thead>
<tbody>
<tr>
<td>Temperature (ºC)</td>
<td>30</td>
<td>25</td>
<td>25</td>
</tr>
<tr>
<td>Salinity (‰)</td>
<td>n.a.</td>
<td>n.a.</td>
<td>&lt;40</td>
</tr>
<tr>
<td>TSS (mg L(^{-1} ))</td>
<td>100</td>
<td>n.a.</td>
<td>50</td>
</tr>
<tr>
<td>pH</td>
<td>6.0 - 9.0</td>
<td>6.5 - 9.0</td>
<td>5.5 - 9.0</td>
</tr>
<tr>
<td>COD (mg L(^{-1} ))</td>
<td>200</td>
<td>30</td>
<td>40</td>
</tr>
<tr>
<td>Oil-Grease (mg L(^{-1} ))</td>
<td>10-20</td>
<td>0.05 - 1</td>
<td>0.01 - 1</td>
</tr>
</tbody>
</table>

n.a.: Not available.

\(^{1}\)Water Pollution Control Regulation of Republic of Turkey (Limit values for Solid Waste Assessment and Disposal Facilities).

\(^{2}\)European Union (Surface Water Quality Standards).

\(^{3}\)Environmental Protection Agency (Parameters of Water Quality - Interpretation and Standards; Recommended or Mandatory Limit Values).
station were lower than the control station (Table 1), suggesting a possible oil input to the system although the oil-grease values were below the limit values (Table 3). Furthermore, COD hasn’t been associated with the isotopic ratios of primary consumers or POM in recently conducted studies where there is a possible oil input to the system. Indeed, no correlations were found between COD and stable isotope values of both zooplankton and POM in the current study. The variation in chl-α and TSS values determined in the region showed that the load in the system may vary for many reasons. For example, the significant linear relationship between TSS and chl-α in the control station can be an indication that TSS in this station may have originated from phytoplankton [32]. In contrast, there was no significant linear relationship between TSS and chl-α at the harbor station, meaning that TSS in the port was not of phytoplanktonic origin but had a different source (Table 2). Furthermore, the salinity values at the control station were significantly higher than the harbor station all through the sampling period (Table 1; p<0.05), indicating that this station can be exposed to the currents flowing from the Mediterranean rather than the ones from the Black Sea. In other words, since the food sources in the control station were of Mediterranean origin, stable isotope ratios may differ from the harbor station. A similar situation has also been determined by [32] and it has been stated that current systems may have an important role in the exchange of food resources.

It is known that terrestrial sources such as detritus and anthropogenic inputs can affect primary consumers and POM's stable isotope compositions [27, 33]. Thus, Fig. 3 shows that POM in the control station (PC) seems to be under the influence of higher carbon sources. Therefore, it can be stated that in this study POM was under the influence of very high carbon-derived phytoplanktonic organisms rather than low-carbon terrestrial inputs (detritus, anthropogenic inputs) in the control station. The chl-α values, indicative of the phytoplanktonic load, were significantly higher in the control station when compared to the harbor station (Table 1; p<0.05) supporting this idea. In addition, the stable C and N ratios of the POM obtained at the control station were similar to the previously conducted studies offshore – far from terrestrial inputs [34, 35]. Therefore, this may support the idea that the control station is not under the influence of terrestrial inputs.

Although there was no significant difference between the stations in terms of δ^15N_{zoo} values, the δ^15N_{zoo} values measured in the harbor were higher than the control station. It is known that δ^15N_{zoo} is generally enriched in coastal regions where anthropogenic terrestrial inputs are high [36,37]. One reason for the enrichment of δ^15N_{zoo} in this region can be the fact that the harbor is closer to the terrestrial sources such as anthropogenic inputs. Similar to the present research, in a study conducted by [38], it has been observed that the δ^15N_{zoo} values were enriched in coastal regions compared to the open seas. Therefore, it can be stated that stable isotope compositions belonging to the zooplankton and POM in the harbor station can be under the influence of terrestrial inputs rather than harbor wastes. Thus, the significant positive correlation between TSS and δ^15N_{zoo} as well as δ^15N_{POM} in the harbor station may support this idea (Table 2).

With this study, it was clearly understood that the load in the harbor station originated from the terrestrial inputs, while the load in the control station originated from phytoplankton, which could affect the isotopic values of zooplankton and POM. In addition, the isotopic values of zooplankton and POM measured during the study were similar to the previous studies.
conducted close to the region and no high differences were observed [11, 39, 40]. Nevertheless, the values of physico-chemical and biological variables were similar to those recorded in the previous studies in the region [2, 4, 5, 41].

Conclusions

Considering the data evaluation in the study area, it can be specified that the wastewater plant discharge in the harbor had a trivial effect on the ecosystem of the coastal area and that the changes in δ13C and δ15N values of the primary consumers and POMs in the region were mostly due to terrestrial inputs, different phytoplanktonic loads and the special current system in the Strait. Additionally, it can be stated that the control station contains food sources of Mediterranean origin while the harbor region contains food sources of Black Sea origin. Thus, being the first stable isotope study on zooplankton and POM in the region, this study is expected to be a reference for future food web studies in the region. However, in order to understand the dynamics better, it is recommended to carry out more comprehensive studies including the effects of estuarine inputs to the strait and other point sources such as sewage inputs.

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Conflict of Interest

The authors declare no conflict of interest.

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