

Short Communication

Isolation and Identification of a Phyllospheric Yeast Strain Capable of Utilizing BDE-3 as Sole Carbon and Energy Source

Xinwei Zhao^{1,2,3}, Yaxiao Wang^{1,2}, Yuling Li^{1,2}, Kejiu Du^{1,2}

¹College of Forestry, Hebei Agricultural University, Baoding 071000, China

²Hebei Provincial Key Laboratory of Tree Species Germplasm Resources and Forest Protection, Baoding 071000, China

³State Key Laboratory of Tree Genetics and Breeding, Chinese Academy of Forestry, Beijing

Received: 9 July 2020

Accepted: 18 August 2020

Abstract

As a class of synthetic organic compounds widely used in varieties of industrial products, polybrominated diphenyl ethers (PBDEs) have attracted increasing attention for the properties of toxicity, bioaccumulation and persistence. In this paper, the phyllospheric yeast strain capable of aerobic degradation of 4-brominated diphenyl ether (BDE-3) was originally isolated from *Populus tomentosa*. Based on the morphological properties and the ITS sequence analysis, the obtained yeast was identified as *Wickerhamomyces anomalus* and designated Y1. The endurance ability of Y1 to BDE-3 exposure was determined by BDE-3 concentration gradient experiment. BDE-3 removal efficiency analysis was conducted by applying GC- μ ECD with initial BDE-3 content of 40 mg/L. Results showed Y1 could remove 83.58 \pm 2.60% of 40 mg/L BDE-3 in 3-day aerobic incubation. The tolerance range of pH and NaCl concentration were determined for further engineering application. It could be the first report about PBDEs degradation by yeast strain, isolated from the phyllosphere of *P. tomentosa* at the non-contaminated site of a subaerial ecological niche, and might be advantageous in comprehensive exploitation of phytoremediation.

Keywords: *Wickerhamomyces anomalus*, BDE-3, phyllosphere, aerobic degradation

Introduction

Applied as one of the most effective and economical brominated flame retardants (BFRs), PBDEs have been widely used for more than three decades in varieties of manufactured materials [1]. Thanks to the extraordinary fire retarding capability, a great deal of wealth and

life has been saved. Since BFRs were first detected in environmental samples in Sweden, interests as well as concerns in environmental distribution, behavior and fate of BFRs including PBDEs have grown sharply [2]. Mainly for massive industrial use and no chemical bond between PBDEs and body material, PBDEs are ubiquitously detected in various environmental media and biological samples even human beings [3]. In addition to the wide distribution, the properties of persistence, bioaccumulation and long distance transportation attracted world-wide concerns. Recently,

*e-mail: dukejiu@126.com

PBDEs have been considered genotoxic, mutagenic and carcinogenic especially for low-brominated diphenyl ethers which are generated from high-brominated diphenyl ethers by debromination [4]. All of these posed detrimental threats to wildlife and human health. Therefore, it's necessary to establish effective technological approaches to deal with or reduce the pollution of PBDEs. In this research, BDE-3 was selected as a model congener of PBDEs to study its aerobic biodegradation.

Biodegradation is one of the most important natural attenuation processes for xenobiotic and synthetic chemicals in environment and deemed as a safe and effective way to remove them from the environment. Many efforts have been taken to degrade PBDEs by applying biotechnological approaches. Among them, a number of microorganisms responsible for debromination reactions have been enriched, isolated and identified from sludge and sediment [5-10]. Deng et al. (2011) investigated *Lysinibacillus fusiformis* DB-1, an aerobic bacteria isolated from sediments, to degrade decabromodiphenyl ether (BDE-209) [6]. Shi et al. (2013) demonstrated the possible aerobic transformation mechanism of PBDE-209 by analyzing of metabolites of BDE-209 metabolized by a *Pseudomonas aeruginosa* bacterium which was isolated from an e-waste dismantling area [7]. Moreover, one microalgae isolate *Chlorella* SICH, isolated from wastewater, could reach 82-90% removal rate of a mixture of BDE-71 and BDE-209 (5:1) at 0.6 mg/L in seven days [11]. In addition to unicellular organisms, plants have been applied to remove PBDEs *in vitro* and *in vivo* [12-14]. Sun et al. (2013) used pumpkin to hydroponically expose to 2,2',4,4'-tetrabromodiphenyl ether (BDE-47) and found the formation of 5 kinds of debromination products and one methoxylated debromination product [13]. Also, Wang et al. (2012) observed the debrominated, hydroxylated and methoxylated products in maize hydroponically exposed to BDE-15, -28 and -47 [12]. Compared with bacteria and plants, little is known about the degradation of PBDEs by yeast strains though the unicellular eukaryotes play important roles in soil ecosystems as well as contaminated sludge and wastewater.

Widely distributed in the environment, various yeast strains have been reported for characterizing the removal capability to some hazardous chemicals. *Magnusiomyces ingens* LH-F1, obtained from the sea mud of a harbor adjacent to an azo dyes factory, was found capable of decolorizing various azo dyes [15]. *Pichia anomala*, isolated from crude oil contaminated soil, could degrade four kinds of polycyclic aromatic hydrocarbons (PAHs) alone or in combination [16]. Moreover, diverse yeast strains have been found participating in industrial wastewater purification with concomitant xenobiotic like color substances [17-19], phenol- and chlorophenol-related substances [19-20]. Because of the structure similarity of aromatic

chemicals to PBDEs, the potential utilization of yeast strains for PBDEs removal seems attractive.

In this study, the yeast strain capable of aerobic degradation of 4-brominated diphenyl ether (BDE-3) was isolated from the phyllosphere of *P. tomentosa*, and identified based on morphological properties and the ITS sequence analysis. The pH and NaCl tolerance capability, growth curves with BDE-3 concentration gradient and BDE-3 removal and assimilation efficiency were investigated.

Material and Methods

Reagents

BDE-3 (purity>98 %) used in this study was purchased from Tokyo Kasei Kogyo Co., Ltd. (TCL) and diluted with alcohol to 80 g L⁻¹ before use. Biochemical reagents and solutions were of analytical grade or higher.

Synthetic Bushnell Hass Mineral Salts (BHMS) medium contains (g/L): KH₂PO₄, 1; K₂HPO₄, 0.2; MgSO₄·7H₂O, 0.2; CaCl₂, 0.02; NH₄NO₃, 1; NaCl, 2; and 2 droplets of 60% FeCl₃. BHMS medium supplemented with desired concentration of BDE-3 was used as BDE-3 degradation medium.

Potato Dextrose Agar (PDA) medium contains (g/L): Potato extract 3, glucose 20, agar 15. PDA medium was for culturing and screening of phenotypically different yeast strains.

Liquid PDA medium containing (g/L): potato extract 3 and glucose 20 was used for activating microorganisms before the growth curve establishment and BDE-3 removal determination analysis. All the culture media were adjusted to pH 6.5 and autoclaved at 115°C for 15 min.

Culturable Yeast Strains Isolation

The *P. tomentosa* forest, aged 16, grows on West Campus of Agricultural University of Hebei, Baoding, China. The leaf material preparation and strains isolation were conducted following the method described by Zhao et al. [21]. The phyllospheric microbiota was subcultured for several times to obtain phenotypically different yeast colonies as many as possible.

Screening BDE-3 Utilization Yeast Strains

Prior to the screening, obtained culturable yeast strains need to be activated by incubating in liquid PDA medium. The initial yeast culture turbidity in liquid PDA was modulated to the optical density at 660 nm (OD₆₆₀) of 0.100, and the suspension culture was incubated on a rotary shaker at 30°C and 150 rpm for 24 h. Cells were harvested by centrifugation (3000 rpm, 10 min, 4°C), and then washed with BHMS medium (3000 rpm, 10 min, 4°C) three times. The pellet of

each strain was resuspended in BHMS medium and used for further investigations including the growth curve establishment and removal efficiency determination. 50 mL activated suspension containing 20 mg L⁻¹ BDE-3 was incubated on a rotary shaker at 30°C and 200 rpm. The OD₆₆₀ was measured every 12h for the establishment of their growth curves using UV-visible spectrophotometer. The strains capable of growing on BDE-3 as sole source of carbon and energy were selected *via* screening 5 days later.

Strain Identification

The morphological properties were investigated to taxonomically characterize the BDE-3 degrading yeast strains. For the molecular identification, Internal Transcribed Spacer (ITS) sequences of obtained yeast were amplified by colony PCR. PCR amplification was performed using the general ITS primers 1 and 4 [22]. The sequence (5'-3') of primer 1 and 4 were TCCGTAGGTGAACCTGCGG and TCCTCCGCTTATTGATATGC, respectively. The amplification reaction was performed in a 50 µl reaction system containing 1×PCR buffer, 0.2 mM dNTPs, 1 µM each of forward and reverse primer, 0.4 mM MgCl₂ and 2U Tag DNA polymerase. The temperature profile was set as follows: on cycle of 5 min at 95°C, followed by 35 cycles of 45 sec at 95°C, 45 sec at 52°C, and 60 sec at 72°C, and one cycle of 10 min at 72°C. The amplification products were subjected to electrophoresis in 1% agarose gel. Sequencing was performed after ITS DNA fragments extraction. The obtained ITS sequences were compared and aligned with sequences deposited in GenBank database using the BLAST program to compare similarity to other yeast ITS sequences. The ITS sequence of degrading yeasts and related sequences obtained from GenBank database were aligned by Clustal X (1.8). The aligned data was used to construct a phylogenetic tree using Neighbor-joining method by MEGA (Version 5.1) with 1000 bootstrap replicates.

Growth Curve Establishment

Growth curve of activated BDE-3 degrading yeast strain was conducted in 150 ml Erlenmeyer flask containing 30 ml BHMS medium supplied with different concentrations of BDE-3 (0, 100, 200, 300, 400, 500 mg L⁻¹). The treatment group without adding BDE-3 was set as blank control. Treatments of BHMS medium supplied with different concentrations of BDE-3 and the one of the same volume of liquid PDA medium were set to analyze the effects of nutritional condition on the yeast growth. The growth curves of all treatment groups were established by OD₆₆₀ measurement. The incubation condition was on a rotary shaker at 30°C and 200 rpm. After every OD₆₆₀ measurement, the batch of used suspensions were removed. Every treatment was set triplicate.

Environmental Factors Tolerance Range Determination

Activated BDE-3 degrading yeast strains cultured in BHMS with initial suspension OD₆₆₀ of 0.100 and augmented with 300 mg L⁻¹ BDE-3 were conducted to determine the tolerance range of pH and NaCl. The range of pH was set as 4.0, 5.0, 6.0, 7.0, 8.0, 9.0, 10.0 and the NaCl concentration (mass concentration) was set as 0, 0.5, 1, 2, 3, 4, 5%. Every treatment was set triplicate. The incubation condition was same to the section 2.5. The OD₆₆₀ of suspension was measured after 5 days incubation.

BDE-3 Removal Efficiency Determination

Activated 30 mL yeast strain BHMS suspension in 150 mL Erlenmeyer flask supplied with 40 mg L⁻¹ BDE-3 was set as removal treatment group. The thermal inactivation treatment group of 30 mL yeast strain BHMS suspension added 40 mg L⁻¹ BDE-3 was set to determine the assimilation and adsorption effects of inactive yeast cells. The BHMS medium supplied with 40 mg L⁻¹ BDE-3 was set as negative control to investigate the potential volatilization or non-biological degradation capacity. Every treatment was biological triplicate. The incubation condition was same to the section 2.5. After 72 h incubation, 10 mL n-hexane was added to incubation suspension. Incubation suspensions were treated on a rotary shaker at 20°C and 150 rpm for 1 h. Then, 1 mL supernate was extracted from the suspensions for removal efficiency analysis by applying gas chromatograph (Agilent Technologies 6890N) coupled with a micro-electron capture detector (Agilent Technologies, USA; GC-µECD). The capillary column, thermal settings and other procedure details were performed following the method described by Zhao et al. [21].

Statistical Analysis

Origin 8.5 was performed to form figures as well as calculate the mean value and standard deviation (SD) for replicates.

Results and Discussion

Culturable Strains Isolation, Screening and Identification

After several times of leaves sampling and strains isolation, a variety of culturable microorganisms on PDA medium were obtained. Through screening, one yeast strain with capability of aerobically metabolizing BDE-3 as sole carbon and energy source was acquired. The yeast colony was mucoid creamy whitish with a diameter of approximate 4 µm on PDA medium (Fig. 1). Yeast cells showed circular with an entire



Fig. 1. The colony morphology of isolated 4-BDE degrading yeast strain.

margin. The molecular identification of the yeast strain was conducted by amplifying and sequencing the ITS fragment. Through aligning with sequences deposited in GenBank database, the obtained yeast showed highest similarity to *Wickerhamomyces anomalus* strain. Based on the morphological properties and the ITS sequence analysis, the isolated BDE-3 degrading yeast strain was identified as *W. anomalus*, and designated Y1. The phylogenetic tree was illustrated in Fig. 2. The accession number of Y1 ITS sequence is MG757836.

Growth Curve Establishment

Activated yeast Y1 suspension supplied with BDE-3 (concentration gradient of 0, 100, 200, 300, 400, 500 mg L⁻¹) was set to determine the endurance capability with initial optical density at 660 nm (OD₆₆₀) of 0.100. As shown in the Fig. 3, the suspension OD₆₆₀ of

blank control group declined over time which illustrated no significant nutrition was residual after activating and washing procedure. The suspension OD₆₆₀ increased along with the BDE-3 concentration from 100 to 500 mg L⁻¹, revealing high BDE-3 concentration up to 500 mg L⁻¹ did not limit the quantitative growth of strain Y1. The highest optical density at 660 nm of most treatment groups (4/5) reached at the second day, then slowly inclined. The highest OD₆₆₀ of 500 mg/L BDE-3 treatment group came up to 0.450 at the 2nd day. The nutrition level effect of liquid PDA medium and BDE-3 in BHMS on yeast biomass was conducted and it could be seen a huge suspension density difference between the two media (Fig. 3), where Y1 in liquid PDA reached the summit of suspension OD₆₆₀ at 3rd day and almost came up to 2.000.

Environmental Factors Tolerance Range Determination

As shown in Fig. 4 (a-b), BDE-3 degrading yeast Y1 could grow on 300 mg L⁻¹ BDE-3 in BHMS medium in pH range of 4.0 to 8.0 and up to 3% NaCl, respectively. If zoom in, the results showed that the yeast Y1 cells grew better at pH 6.0 than 7.0, illustrating the yeast was mildly acidophilic. And yeast Y1 grew best in BHMS medium when no extra NaCl was added but could still survive at 3% NaCl, proving the yeast possessed strong salt tolerance capability.

Removal Efficiency Determination

The removal efficiency of BDE-3 degrading yeast Y1 was conducted by aerobically removing 40 mg L⁻¹ BDE-3 with initial suspension OD₆₆₀ of 0.100. The results showed that no significant BDE-3 loss appeared in negative control group, thus ruling out the

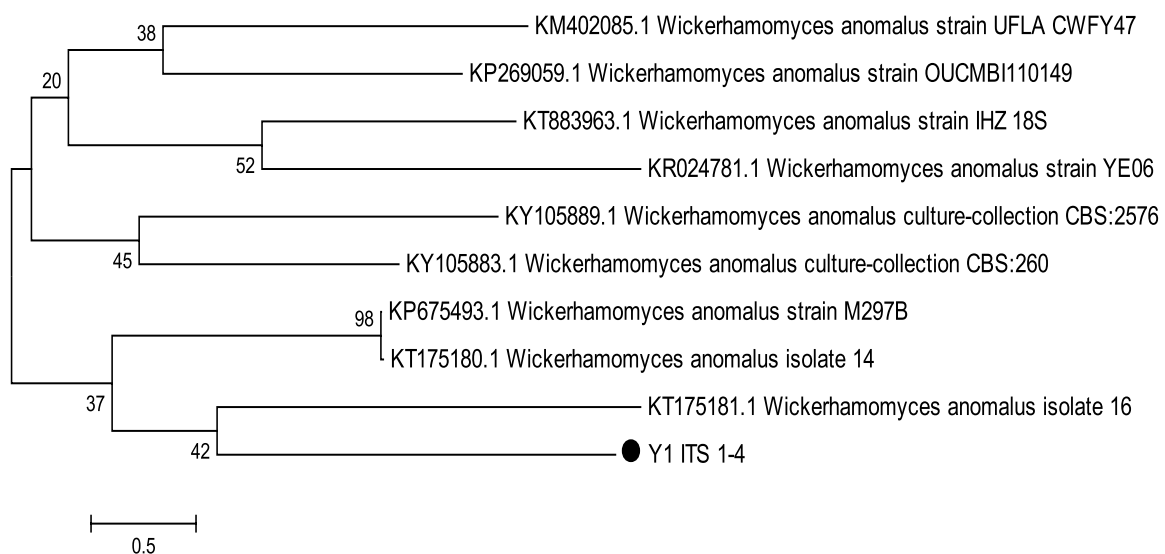


Fig. 2. The phylogenetic tree of yeast Y1 based on ITS sequence. The black point represents the isolated phyllosporic yeast Y1, the 9 alignment strains are shown by accession number and name of species.

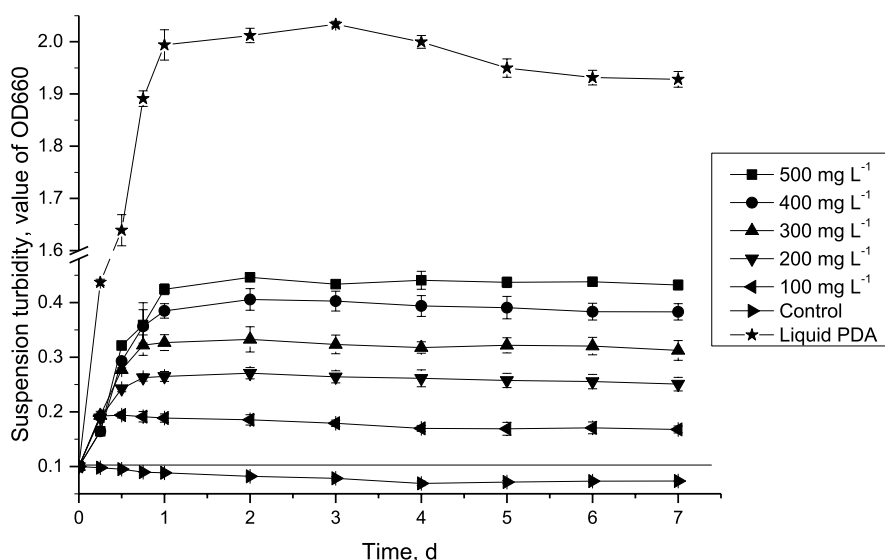


Fig. 3. Y1 growth curves in 4-BDE BHMS and liquid PDA media. The baseline of 0.1 means the initial suspension turbidity. The character d means day. The baseline of 0.1 means the initial suspension turbidity.

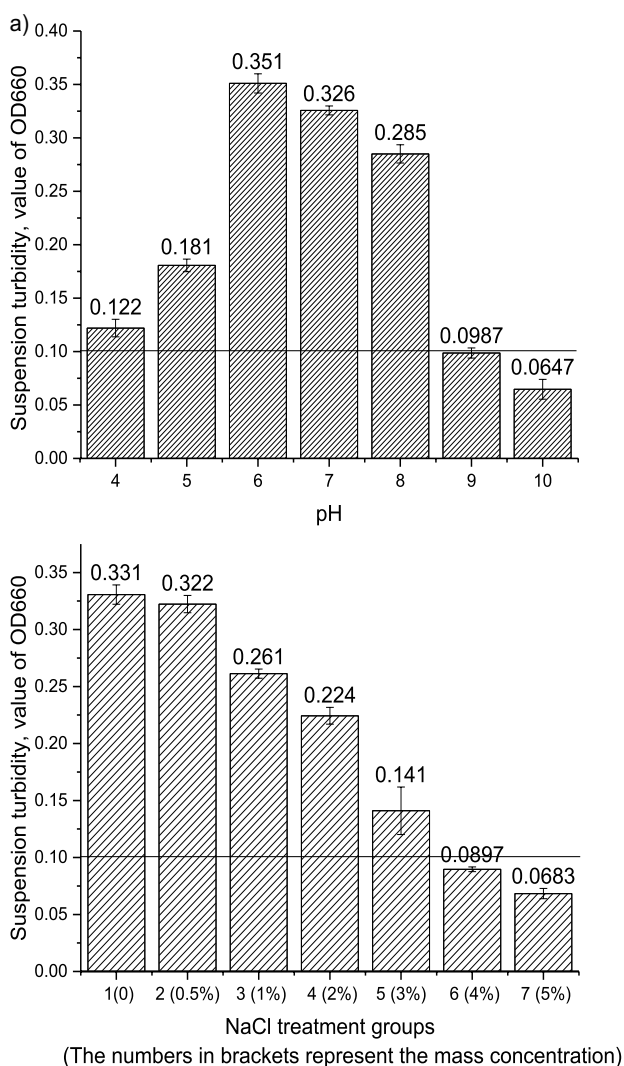


Fig. 4. Yeast Y1 suspension turbidity in the range of pH a) and NaCl concentration b) after 5d incubation. The baseline of 0.10 means the initial suspension turbidity

possibility of significant volatilization or non-biological degradation of BDE-3. For biological removal treatment group, Y1 could transform 83.58 ± 2.60 % of 40 mg L^{-1} BDE-3 in 3-day incubation. For thermally inactivated Y1 suspension, inactive Y1 cells still could absorb and assimilate 41.13 ± 4.95 % of 40 mg L^{-1} BDE-3 (Table. 1).

The BDE-3 degrading yeast Y1 was originally isolated from the phyllosphere of *P. tomentosa* forest, where the survival environment is not significantly or continuously contaminated by PBDEs or other organic pollutants. For acquiring microbial resource for potential application to degrading hazardous materials, the contaminated areas of corresponding pollutant were the preferred investigation sites, where these sites could domesticate microorganisms with outstanding adaptive mechanisms and degrading properties. Most researches involved acquiring desired PBDEs-degrading strains were focused on sites of PBDEs contaminated field [8, 23, 24]. The finding of yeast Y1, with the specific ecological niche on phyllosphere, broadens the existing range of PBDEs degrading microorganisms, and as unicellular fungi, strain Y1 is the first report with property of biodegrading PBDEs. Phyllosphere, the aerial above-ground plant part, is a high-density microbial attachment area where more versatile noxious-substance degrading microorganisms to be further explored [25-26].

The addition of carbon and energy source to culture environment was needed necessarily for partial PBDEs degrading strains to maintain efficient removal rate [7, 27]. Moreover, the addition of extra energy materials could be beneficial to promoting cell surface hydrophobicity, thus enhancing the removal efficiency [23, 24, 28]. In the present research, the obtained yeast Y1 could grow on BDE-3 as sole carbon and energy source, which is great favor for engineering application: the utilization of yeast Y1 could reduce the economical

Table 1. The BDE-3 removal efficiency of different treatment groups.

Treatments	Removal efficiency (%)
BDE-3 removal group ^a	83.58±2.60
BDE-3 absorb group ^b	41.13±4.95
Blank control group ^c	0.30±0.14

a: Y1 suspension with supplied of 40 mg L⁻¹ BDE-3.

b: Thermally inactivated Y1 suspension with supplied of 40 mg L⁻¹ BDE-3. c: BHMS medium without Y1 cells but 40 mg L⁻¹ BDE-3. The removal efficiency is presented by means±SD of triplicates

and environmental burden caused by adding artificial co-metabolites or surfactants, and thus eliminate the possible adverse consequence if the additive materials are poisonous like toluene or other compounds possessing benzene ring structure.

In this paper, the aerobic transformation efficiency of yeast Y1 reached 83.58±2.60 % of 40 mg L⁻¹ BDE-3 in 3 days incubation, which is fairly or even more effective when comparing with the previous reports. For instance, *Rhodococcus* sp. RR1 aerobically transformed 20% of 17 µg L⁻¹ BDE-3 within 72 hours [29]. *Pseudomonas putida* degraded 42.5% of 2 mg L⁻¹ BDE-3 after 10 days incubation [10]. Two kinds of anaerobic sludge removed 77% and 31% of 5 mg L⁻¹ BDE-3 respectively over 16 days [30]. *Sphingomonas* sp. PH-07 could degrade 23% of 1 g L⁻¹ BDE-3 within 8 days [5].

Conclusions

The yeast strain with the ability to degrade 4-brominated diphenyl ether (BDE-3) was isolated from the phyllosphere of *P. tomentosa* at a subaerial ecological niche without any contamination of synthetic chemicals. The strain was identified as *Wickerhamomyces anomalus* and designated Y1 whose pH tolerance rang was 4.0 to 8.0, and could grow in the BHSM medium adding 3% NaCl. The yeast Y1 could remove 83.58±2.60 % of 40 mg L⁻¹ BDE-3 under an aerobic incubation condition within 3 days. The findings are expected to inspire the integral phytoremediation combining plant and associated above-ground microorganisms for gaseous PBDEs removal.

Acknowledgements

This work was supported by the Natural Science Foundation of Hebei Province (C2018204134), the National High Technology Research and Development Program of China (2012AA101403), and the National Natural Science Foundation of China (30972384).

Contributions

Kejiu Du contributed to the experimental design; Xinwei Zhao, Yaxiao Wang, Yuling Li, contributed significantly to analysis and manuscript preparation; Xinwei Zhao, Kejiu Du performed the data analyses and wrote the manuscript.

Conflict of Interest

The authors declare no conflict of interest.

References

1. SONG W., FORD J.C., LI A., STURCHIO N.C., ROCKNE K.J., BUCKLEY D.R., MILLS W.J. Polybrominated diphenyl ethers in the sediments of the Great Lakes. 3. Lakes Ontario and Erie. *Environ. Sci. Technol*, **39**, 5600, **2005**.
2. ANDERSSON Ö., BLOMKVIST G. Polybrominated aromatic pollutants found in fish from Sweden. *Chemosphere*, **10**, 1051, **1981**.
3. LAW R.J., COVACI A., HARRAD S., HERZKE D., ABDALLAH M.A., FERNIE K., TOMS L.M., TAKIGAMI H. Levels and trends of PBDEs and HBCDs in the global environment: status at the end of 2012. *Environ Int*, **65** (2), 147, **2014**.
4. RICHARDSON V., STASKAL D., ROSS D., DILIBERTO J., DEVITO M., BIRNBAUM L. Possible mechanisms of thyroid hormone disruption in mice by BDE 47, a major polybrominated diphenyl ether congener. *Toxicol. Appl. Pharm*, **226**, 244, **2008**.
5. KIM Y.M., NAM I.H., MURUGESAN K., SCHMIDT S., CROWLEY D.E., CHANG Y.S. Biodegradation of diphenyl ether and transformation of selected brominated congeners by *Sphingomonas*, sp. PH-07. *Appl Microbiol Biotechnol*, **77** (1), 187, **2007**.
6. DENG D., GUO J., SUN G., CHEN X.J., QIU M.D., XU M.Y. Aerobic debromination of deca-BDE: Isolation and characterization of an indigenous isolate from a PBDE contaminated sediment. *Int Biodeterior Biodegradation*, **65** (3), 465, **2011**.
7. SHI G., YIN H., YE J., PENG H., LI J., LUO C. Aerobic biotransformation of decabromodiphenyl ether (PBDE-209) by *Pseudomonas aeruginosa*. *Chemosphere*, **93** (8), 1487, **2013**.
8. ZHANG S.W., XIA X.H., XIA N., WU S., GAO F., ZHOU W.W. Identification and biodegradation efficiency of a newly isolated 2,2',4,4'-tetrabromodiphenyl ether (BDE-47) aerobic degrading bacterial strain. *Int Biodeterior Biodegradation*, **76**, 24, **2013**.
9. LI Y., CHEN L., NGOC D.M., DUAN Y.P., LU Z.B., WEN Z.H., MENG X.Z. Polybrominated diphenyl ethers (PBDEs) in PM2.5, PM10, TSP and gas phase in office environment in Shanghai, China: occurrence and human exposure. *Plos One*, **20**, 10(3), e0119144, **2015**.
10. LV Y.C., LI L.H., CHEN Y.C., TANG Z.H., HU Y.Y. Effects of glucose and biphenyl on aerobic cometabolism of polybrominated diphenyl ethers by *Pseudomonas putida*: Kinetics and degradation mechanism. *Int Biodeterior Biodegradation*, **108**, 76, **2016**.

11. DENG D., TAM F.Y. Isolation of microalgae tolerant to polybrominated diphenyl ethers (PBDEs) from wastewater treatment plants and their removal ability. *Bioresour Technol*, **177**, 289, **2015**.
12. WANG S., ZHANG S., HUANG H., LU A., PING H. Debrominated, hydroxylated and methoxylated metabolism in maize (*Zea mays* L.) exposed to lesser polybrominated diphenyl ethers (PBDEs). *Chemosphere*, **89** (11), 1295, **2012**.
13. SUN J., LIU J., YU M., WANG C., SUN Y., ZHANG A., WANG T., LEI Z., JIANG G. In vivo metabolism of 2,2',4,4'-tetrabromodiphenyl ether (BDE-47) in young whole pumpkin plant. *Environ Sci Technol*, **47** (8), 3701, **2013**.
14. YU M., LIU J., WANG T., SUN J., LIU R., JIANG G. Metabolites of 2,4,4'-tribrominated diphenyl ether (BDE-28) in pumpkin after in vivo and in vitro exposure. *Environ Sci Technol*, **47** (23), 13494, **2013**.
15. TAN L., LI H., NING S., XU B. Aerobic decolorization and degradation of azo dyes by suspended growing cells and immobilized cells of a newly isolated yeast *Magnusiomyces ingens* LH-F1. *Bioresour Technol*, **158** (4), 321, **2014**.
16. PAN F., YANG Q., ZHANG Y., ZHANG S., YANG M. Biodegradation of polycyclic aromatic hydrocarbons by *Pichia anomala*. *Biotechnol Lett*, **26** (10), 803, **2004**.
17. MEEHAN C., BANAT I.M., MUMULLAN G., NIGAMP P., SMYTH F., MARCHANT R. Decolorization of Remazol Black-B using a thermotolerant yeast, *Kluyveromyces marxianus* IMB3. *Environ Int*, **26** (1-2), 75, **2000**.
18. WU Y., HU Y., XIE Z., FENG S., LI B., MI X. Characterization of biosorption process of acid orange 7 on waste brewery's yeast. *Appl Biochem Biotech*, **163**, 882, **2011**.
19. AHUATZI-CHACON D., ORDORICA-MORALES G., RUIZ-ORDAZ N., CRISTIANI-URBINA E., JUAREZ-RAMIREZ C., GALINDEZ-MAYER J. Kinetic study of phenol hydroxylase and catechol 1,2-dioxygenase biosynthesis by *Candida tropicalis* cells grown on different phenolic substrates. *World J Microb Biot*, **20**, 695, **2004**.
20. BERGAUER P., FONTEYNE P.A., NOLARD N., SCHINNER F., MARGESIN R. Biodegradation of phenol and phenol-related compounds by psychrophilic and cold-tolerant alpine yeasts. *Chemosphere*, **59**, 909, **2005**.
21. ZHAO X.W., LIU M., LIU Y.C., LI Y.L., ZHUANG S., DU K.J. Aerobic degradation of 4-monobrominated diphenyl ether: identifying and characterizing three strains isolated from *Populus tomentosa* phyllosphere. *Pol. J. Environ. Stud*, **26** (3), 1385, **2017**.
22. WHITE T., BRUNS T., LEE S., TAYLOR J. In: PCR - Protocols and Applications - A Laboratory Manual, San Diego, Academic Press, 315, **1990**.
23. TANG S., YIN H., CHEN S., PENG H., CHANG J., LIU Z. Dang Aerobic degradation of BDE-209 by *Enterococcus casseliflavus*: Isolation, identification and cell changes during degradation process. *J Hazard Mater*, **308**, 335, **2016**.
24. STIBOROVA H., VRKOSLAVOVA J., LOVECKA P., PULKRABOVA J., HRADKOVA P., HAJSLOVA J., DEMNEROVA K. Aerobic biodegradation of selected polybrominated diphenyl ethers (PBDEs) in wastewater sewage sludge. *Chemosphere*, **118**, 315, **2015**.
25. LINDOW S.E., BRANDL M.T. Microbiology of the phyllosphere. *Appl Environ Microbiol*, **69** (4), 1875, **2003**.
26. REDFORD A.J., BOWERS R.M., KNIGHT R., LINHART Y., FIERER N. The ecology of the phyllosphere: geographic and phylogenetic variability in the distribution of bacteria on tree leaves. *Environ Microbiol*, **12** (11), 2885, **2010**.
27. CHEN J., ZHOU H.C., WANG C., ZHU C.Q., TAM N.F. Short-term enhancement effect of nitrogen addition on microbial degradation and plant uptake of polybrominated diphenyl ethers (PBDEs) in contaminated mangrove soil. *J Hazard Mater*, 300:84-92, **2015**.
28. TANG S., BAI J., YIN H., YE J., PENG H., LIU Z., DANG Z. Tea saponin enhanced biodegradation of decabromodiphenyl ether by *Brevibacillus brevis*. *Chemosphere*, **114** (22), 255, **2014**.
29. ROBROCK K.R., COELHAN M., SEDLAK D.L., ALVAEZ-COHEN L. Aerobic biotransformation of polybrominated diphenyl ethers (PBDEs) by bacterial isolates. *Environ Sci Technol*, **43**, 5705, **2009**.
30. SHIH Y.H., CHOU H.L., PENG Y.H. Microbial degradation of 4-brominated diphenyl ether with anaerobic sludge. *J Hazard Mater*, **30**, 213-214:341-6, **2012**.