Ground-Level Ozone Production over an Industrial Cluster of China: a Box Model Analysis of a Severe Photochemical Pollution Episode

Xue Yang¹³, Xiang Cheng², Huaizhong Yan³, Youmin Sun¹, Guiqin Zhang¹

¹College of Municipal and Environmental Engineering, Shandong Jianzhu University, Ji'nan 250101, China
²Shandong Zibo Ecological Environment Monitoring Center, Zi’bo 255040, China
³Shandong Jinan Ecological Environment Monitoring Center, Ji’nan 250101, China

Received: 18 June 2021
Accepted: 21 October 2021

Abstract

To better understand the characteristics and processes of ozone formation over the industrial cluster of China, intensive field measurements of atmospheric O₃, volatile organic compounds (VOCs), and related parameters were conducted in a typical industrialized city of China, Zibo, from July 1 to July 31, 2018. Monitoring data and comparison with other observations revealed severe O₃ pollution in Zibo with high frequency (~55%) of the non-attainment O₃ episodes and high average O₃ levels (~42 ppbv). An observation-constrained chemical box model (OBM) was deployed to dissect the O₃ formation mechanism and ozone-precursor relationship during a severe photochemical smog episode. O₃ was produced by NO + HO₂ and NO + RO₂ reaction and destroyed via the NO₂ + RO₂ and NO₂ + OH reaction. The observed O₃ pollution was dominated by intense in-situ O₃ chemical formation with regional transport playing a negative contribution. Sensitivity studies suggested that in situ O₃ production was in a VOC-limited or mixed-control regime with aromatics (mainly 1,3,5-Trimethylbenzene and toluene) being the dominant parent hydrocarbons of O₃. The OBM and positive matrix factorization (PMF) results indicated that the most efficient way to alleviate the O₃ pollution in Zibo is to reduce emissions of aromatics from solvent use, industrial processes and vehicles.

Keywords: ozone, industrial VOCs, observation-based model, in situ photochemical formation, aromatics

*e-mail: zhangguiqin320@163.com
Introduction

Atmospheric pollution has been sharply increasing over the past three decades due to rapid urbanization and industrialization [1,2]. In particular, chronic ozone (O₃) exceedances are considered to be the main air pollution problem in China in recent years [3-7]. O₃ is a key oxidant and greenhouse gas, and its photolysis leads to hydroxyl radical (OH) formation. O₃ also has a detrimental impact on human well-being, vegetation, and regional air quality [8-11]. Therefore, mitigation of O₃ pollution in China is important due to its significant role in air quality and climate change.

In principle, photochemical reactions between volatile organic compounds (VOCs) and nitrogen oxides (NOₓ = NO + NO₂) control the O₃ budget in the low troposphere [12, 13]. Challenges in mitigation of O₃ pollution mainly lie in understanding the non-linear relationship of ozone precursors [14]. As key precursors of O₃, VOCs are composed of hundreds of species and have a variety of sources and adverse health effects [15, 16]. Ambient VOCs are emitted directly from various sources, including anthropogenic activities (e.g., motor vehicles and industrial processes) and natural and biomass burning sources [17-19]. In recent years, rapid industrialization in China was accompanied by large emissions of industrial VOCs. Cai et al. [20] reported a considerable contribution of industrial VOCs (~36%) to ambient VOCs in Shanghai and An et al. [21] found that industrial emissions contributed 45-63% of total VOCs in Yangtze River Delta (YRD). Therefore, it is urgent to identify the characteristics of industrial VOCs and their contribution to O₃ formation from a regional perspective.

Characterization of VOCs and their effect on O₃ has long been carried out in several regions around China [22, 23]. Knowledge of the ozone production regime (VOC-limited vs. NOₓ-limited) is critical to understanding in-situ ozone production. In the past decade, the observed based model (OBM) based on Master Chemical Mechanism (MCM), Carbon bond IV mechanism (CB5) or Regional Atmospheric Chemical Mechanism (RACM) has been widely used to further understand the O₃ photochemistry [4, 24-28]. For instance, using observed based model (OBM), Kanaya et al. [25] examined the net photochemical production rate of O₃ at Mount Tai (36.25°N, 117.10°E, 1534 m above sea level (ma.s.l.)), a typical mountain site over North China Plain (NCP) region and found that the O₃ production regime was typically NOₓ-limited. Cheng et al. [28] indicated that the O₃ production in Guangzhou and Hong Kong was VOC-limited and further selected 12 compounds, i.e., formaldehyde, acetaldehyde, m,p-xylene, toluene, etc., which contributed to 85-89% O₃ formation.

Xue et al. [26] analyzed the relationship of ozone (O₃) and its precursors in four large Chinese cities, namely, Beijing, Shanghai, Guangzhou and Lanzhou, and confirmed that O₃ production was in a VOC-limited regime in Beijing, Shanghai and Guangzhou, and a NOX-limited regime in Lanzhou. Using the MCM simulation, Lyu et al. [29] revealed that O₃ production was generally VOC-limited in Wuhan, and that ethene and toluene were the primary contributors to O₃ formation. Most previous studies have focused on VOCs and O₃ pollution in urban and/or remote areas, while analyses of VOC and O₃ pollution in industrial areas are sparse. In addition, there are some uncertainties in RACM and CB because of the lumping of VOC components having similar chemical structures and activity. Thus, using a model equipped with the Master Chemical Mechanism (MCM version 3.3; http://mcm.leeds.ac.uk/MCM/) is likely most appropriate for simulating in-situ O₃ production and exploring the relationships of ozone precursors over industrial cluster in China.

Zibo is one of the most important industrial production bases of Shandong Province in China and is also the main transportation city in the Beijing-Tianjin-Hebei pollution transport channel. It was estimated by the Department of the Ecological Environment of Shandong Province that in 2015, anthropogenic emissions of VOCs were 0.02 Tg in Zibo, ranking third in Shandong Province. Such high emissions over a relatively small land area lead to much higher VOCs emission densities in Zibo compared with other regions (i.e., YRD, 2.77 Tg in 2007 [30]). The objective of this research was to better understand O₃ photochemistry in Zibo from July 1 to 31, 2018, through the application of the observation-based model (OBM). In the following discussion, we first provide the temporal variations of O₃ formation and the potential impact to downstream regions over the industrial cluster of China, which may have general implications for understanding the causes of regional secondary pollution in China.

Material and Methods

Site Description

The measurement site is an industrial site located in the northern suburb of Zibo (36°52′21″N-117°58′54″E), which is surrounded by industrial parks and traffic (see Fig. 1). As shown in Fig. 1, the Zhangdian industrial area is located approximately 1.1 km southwest of the sampling site. Several industries have a presence in this area, including the chemical, architecture, electronics, and paint industries. The Zhoucun industry area is about 8.9 km away from the sampling site. This area contains several pollution sources, including pharmaceutical and chemical industries and the printing industry. As the prevailing winds in summer were
from the south/southeast under the influence of the Asian monsoon, the study site is downwind of the above two industrial areas.

Instruments and Data

\( \text{O}_3 \), \( \text{O}_3 \) precursors as well as meteorological parameters were continuously measured from July 1 to July 31, 2018 by a set of techniques after strict quality control. In brief, \( \text{O}_3 \) was measured by a commercial ultraviolet absorption analyser (Thermo 49i, US); \( \text{NO}_X \) with a chemiluminescence analyser with a molybdenum oxide converter (Thermo 42i, US); CO by a non-dispersive infrared analyser (Thermo 48i, US). Meteorological parameters including temperature, relative humidity (RH), wind speed and direction were continuously recorded by an automatic weather station (MILOS520, Vaisala, Finland). 55 VOC species, including 29 alkanes, 9 alkenes, 16 aromatics and 1 alkyne (ethyne) were measured and quantified by commercial automatic GC5000 analysis systems (see Table 1). The VOCs measurements were conducted online throughout the observation period with the time resolution of 1 h. To ensure the accuracy of test data, the instrument is calibrated before monitoring with the standard photochemical gas for ozone precursors approved by the US Environmental Protection Agency (EPA) Photochemical Assessment monitoring Station (PAMS). Five-level calibration curves were used to calibrate the equipment and quantify the VOC concentrations during the whole campaign. During the calibration, the correlation coefficient was between 0.992 and 0.999. The detection limits of those measured VOCs ranged from 3 to 50 pptv. The detailed description of quality assurance and control procedures of all the above instruments has been described in many previous studies (e.g., An et al. [21] and Zou et al. [31]).

Observation-Based Model (OBM)

The Observation-based model (OBM) based on the newest version of the MCM is a robust tool for explicit simulation of atmospheric chemistry and describes ~17000 chemical reactions of individual VOCs [32, 33]. The model is a zero-dimensional chemical box model which includes homogeneous chemistry, heterogeneous chemical processes, dry deposition, and dilution mixing within the boundary layer and has been successfully applied to simulate \( \text{O}_3 \) formation, PAN production, atmospheric oxidative capacity, ROx chemistry and carbonyls secondary formation [34-37]. Specifically, dry deposition of gases and VOCs was adopted in our model with dry deposition velocities taken form Zhang et al. [38]. In addition, the dilution mixing within the boundary layer for all species was applied in the model with the boundary layer height being assumed to be 300 m at night and linearly rise to 1500 m in the early afternoon. Sensitivity tests by changing maximum PBL heights (e.g., 1000 m and 2000 m) has been conducted in our previous study [36], which indicated negligible impacts (i.e., <3% to net carbonyls production rates) on the modeling conclusion. Here, we use observed \( \text{O}_3 \) and its precursors as well as related meteorological parameters to simulate the rates and regimes of photochemical \( \text{O}_3 \) production. Specially, the model reads in the observation data every hour to calculate
the in-situ rates of O₃ production and loss. The net O₃ production rate (Net P (O₃)) can be estimated from the P (O₃) and L (O₃). The corresponding rates can be determined from the following equations:

\[
P(O₃) = kₜ[HO₂][NO] + \sum (kₗ[RO][NO])
\]

(1)

\[
L(O₃) = kₚ[O^3(Δ)][H₂O] + \sum (kₚ[OH][H₂O]) + \sum (kₙ[VOC][O₃]) + \sum (kₚ[VOC][NO])
\]

(2)

\[
Net P(O₃) = P(O₃) - L(O₃)
\]

(3)

Where RO₂ is the representative of organic peroxy radicals (e.g. C₇H₈O₂), VOC represents unsaturated NMHC species, k represent related reaction rate coefficients. In addition to gas phase reactions, heterogeneous processes regarding N₂O₅ and NO₃ loss were also included in our model. Note the kinetic data are obtained from the IUPAC database (http://iupac.pole-ether.fr/index.html). In particular, the OBM model simulated O₃ and reactive carbonyls (e.g. HCHO) with constraints of VOC precursors agree well with the observed diurnal patterns by Xue et al. [37] and Yang et al. [36], indicating their dominate secondary source and suggesting the OBM is capable of simulating the concentrations of observed highly reactive species.

In detail, the model was constrained by the intensive set of observations to represent the measured real atmospheric conditions. For instance, the observed concentrations of O₃, VOCs, CO, NO, NO₂, SO₂, temperature and H₂O were processed (averaged or interpolated) into the model inputs with a time resolution of 1 hour and used to constrain the model. The model calculations were performed for one smog episode with concentrations of O₃ exceeding the ambient air quality standard (i.e., 2 to 5 July 2018). In the base run, the model calculation started from 00:00 local time and runs for a 24h period. The model was run five times to stabilize the simulation of the unconstrained compounds and the outputs of the last run were subject to further analyses. Detailed information about the model configuration and calculation of the O₃ production rate has been described elsewhere [39].

### Results and Discussion

#### Characteristics of Ozone Pollution

Fig. 2 shows the measured time series of O₃ pollution and meteorological conditions at Zibo during the intensive campaign. Weather conditions consisted of relatively high temperatures (with average±standard deviation of 29.5±3.6°C) and moderate relative humidity (72.6±12.4%). During 2-5 July (four days), Zibo was hit by one prolonged photochemical smog episode during the early stages of sampling period with peak O₃ value exceeding the national ambient air quality standards.

<table>
<thead>
<tr>
<th>Category</th>
<th>Species</th>
</tr>
</thead>
<tbody>
<tr>
<td>Alkanes(29)</td>
<td>2-methylpentane, n-decane, n-hexane, i-butane, 2,2,4-trimethylpentane, 2,3,4-trimethylpentane, n-heptane, n-butane, 2,3,4-trimethylpentane, 3-methylpentane, n-octane, 2,2-dimethylpentane,</td>
</tr>
</tbody>
</table>
Ground-Level Ozone Production...

of China (Class II: ~93 ppb for hourly values), when high temperature and lower RH persisted, which is representative of local Zibo pollution. The zoomed-in figure of this episode is given in Fig. S1. The peak value of \( O_3 \) was recorded at 163 ppbv on 4 July, 2018. In comparison, relatively low \( O_3 \) levels were observed during the middle and later period of the campaign accompanied with relatively lower temperatures (28.5±3.7°C) and higher RH (80.1±9.2%). However, the high NOx and VOCs values (19.60 and 66.27 ppbv) were observed during 25–28 July, 2018, which may be interpreted as the weakened local photochemistry with lower RH levels and temperatures described above. Obviously, the inspection of observational data shows the serious situation of summertime ozone pollution over the industrial cluster of China during the smog episode.

Fig. 3 shows the average diurnal profiles of \( O_3 \), total VOCs (TVOCs), related trace gases, and meteorological parameters at Zibo derived from the measurements. This figure illustrates the well-defined unimodal profile for diurnal variations of \( O_3 \) in the polluted industrial environments. The mean \( O_3 \) mixing ratios were at a minimum of ~21 ppbv at 05:00 LT (local time) and a maximum of ~83 ppbv at 14:00 LT. The broad peak in \( O_3 \) concentrations during the afternoon and the dramatic average daytime \( O_3 \) build-up (i.e., 62 ppbv) indicate the strong in situ ozone production during daytime in the study area. The mean values for NOx showed a unimodal profile with a predominant peak (~28 ppbv) at 06:00 LT. Meanwhile, NOx concentrations were high (~25 ppb) at night, which was likely due to the shallow boundary layer and enhanced truck emissions during the night. In comparison, the NOx values in daytime were low, which may attribute to not only the deeper boundary layer but also the reaction with OH and RO2 as well as self-photolysis. CO exhibited a less significant diurnal pattern (with a mean value in the range of 0.66-0.99 ppmv). In contrast, as shown in Fig. 3, the TVOCs exhibited a bimodal pattern, i.e., morning (08:00-09:00 LT), afternoon maxima (16:00-18:00) and noon minimum (14:00-15:00) due to increased photochemical activity during the daytime. Such results were similar to the pattern observed in a typical petrochemical area of Shanghai [40]. Overall, an inspection of diurnal variations clearly indicates the strong photochemical ozone formation at Zibo.

**Composition of Observed VOCs and Comparison with Other Locations**

As mentioned above, markedly severe \( O_3 \) pollution was observed at Zibo during the sampling period. Individual VOC compounds have different photochemical reactivities and \( O_3 \) formation capacity. Therefore, we examined the VOCs compositions and reactivities obtained at Zibo during the summer campaign in this section to evaluate the contributions of the individual VOCs to ozone production. Concentration levels and photochemical properties i.e., OH reactivity (\( R_{OH} \)) of 55 VOC species measured at Zibo were
calculated and depicted in Fig. 4. As depicted in Fig. 4, during the whole campaign, the average concentrations of alkanes, alkenes, alkynes, and aromatics were 16.29, 2.66, 5.52 and 7.28 ppbv, respectively. The three most abundant VOCs were ethane (5.16 ppbv), ethyne (3.50 ppbv) and m-diethylbenzene (12.36 ppbv), representing 43% of the total VOCs measured. The reactivity of a particular VOCs was calculated as the product of its concentration and its rate constant for reaction with OH: 
\[ R_{\text{OH}} = [\text{VOC}] \times K_{\text{OH}}. \]

The total hydrocarbon reactivity for the study was 6.13 (±0.09) s⁻¹ at Zibo, which is comparable to those measured many urban areas, such as Shanghai (5.85 s⁻¹), Guangzhou (5.23 s⁻¹) and Lanzhou (9.33 s⁻¹) [26], while much higher than Yucheng (2.07 s⁻¹), a rural site downwind of Ji’nan [41]. Furthermore, ozone formation potential (OFP) were also calculated to evaluate the contributions of the individual VOCs to ozone production with an average of 284.67 μg m⁻³, which is much higher than that measured in the industrial area of Brazil [42]. This clearly indicates that the reactive VOCs are quite high at Zibo, most probably due to a large number of industrial sources near the study site. Based on the OH reactivity calculation result, aromatics were the most important reactive VOCs group at Zibo and accounted for 39% of the total hydrocarbon reactivity. For the particular reactive aromatics, the most significant species were 1,3,5-Trimethylbenzene (0.88 s⁻¹) with a mean contribution of 14% on average, of the hydrocarbon reactivity, followed by styrene (0.60 s⁻¹; 10%). Isoprene was the most reactive biogenic hydrocarbon with an average \( R_{\text{OH}} \) value of 0.88 s⁻¹ and contribution of 14%.

For a better understanding of the \( O_3 \) and VOCs pollution situations at Zibo, we compare the \( O_3 \), TVOC concentrations and dominant types of VOCs at Zibo with those measured in other industrial, urban, and suburban locations (see Table 2). The average concentration of TVOCs is within the range reported for other industrial environments. In general, the concentrations of TVOCs at Zibo were higher than those measured in industrial areas of Houston [43], but lower than Nanjing [44] and Lanzhou [45]. Generally, the absolute concentration of VOCs and \( O_3 \) is affected by many factors, e.g., sampling site, time, environmental and meteorological conditions. The higher VOCs and \( O_3 \) concentrations in Lanzhou are uniquely due to its valley topography [45]. What’s more, compared with urban/suburban areas, the VOCs value was lower than those in Beijing [46], Tianjin [47], Guangzhou [31] and Hong Kong [48] but was comparable to the result obtained in Wuhan (see Table 2). In addition, the dominant VOCs file was similar in both industry and urban sites, and the high abundant aromatics suggest that the solvent usage would be important sources at Zibo. Table 2 also shows that the average and maximum \( O_3 \) concentrations at Zibo are generally greater than those reported by
the studies conducted in other urban, industrial, and suburban locations, which highlights the severity of photochemical pollution over Zibo.

The Process Analysis of Ozone Episode

The aforementioned analysis qualitatively revealed the effects of in situ photochemical production to the observed ozone pollution at Zibo. In this section, we further evaluate the chemical budgets of O$_3$ using a detailed case study with the aid of the OBM model. The detailed measurement data of O$_3$ and meteorological parameters in this episode (i.e., 2 to 5 July 2018) are depicted in Fig. S1. During this episode, the atmosphere was characterized by persistent high ozone concentrations (in the range of 111-163 ppbv) and high NOx (~30 ppbv as the daytime average). Meanwhile, the VOCs concentration on 4 July was

![Fig. 4. a) OH reactivity and b) corresponding concentration of individual VOCs (in unit of ppbv).](image)

<table>
<thead>
<tr>
<th>Location</th>
<th>Sites category</th>
<th>Sampling period</th>
<th>TVOCs (ppbv)</th>
<th>Alkanes</th>
<th>Alkenes</th>
<th>Aromatics</th>
<th>O$_3$-average (ppbv)</th>
<th>O$_3$-maximum (ppbv)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Zibo</td>
<td>Industrial</td>
<td>Jul. 2018</td>
<td>31.75</td>
<td>16.29</td>
<td>2.66</td>
<td>7.28</td>
<td>42</td>
<td>163</td>
</tr>
<tr>
<td>Nanjing [44]</td>
<td>Industrial</td>
<td>May-Aug.2013</td>
<td>34.40</td>
<td>14.98</td>
<td>7.35</td>
<td>9.06</td>
<td>32</td>
<td>146</td>
</tr>
<tr>
<td>Houston [43]</td>
<td>Industrial</td>
<td>Aug-Sep. 2006</td>
<td>20.45</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>-</td>
</tr>
<tr>
<td>Lanzhou [45]</td>
<td>Industrial</td>
<td>Feb-Mar. 2017</td>
<td>50.05</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>-</td>
</tr>
<tr>
<td>Wuhan [29]</td>
<td>Urban</td>
<td>Feb 2013-Oct 2014</td>
<td>24.3</td>
<td>13.3</td>
<td>4.3</td>
<td>4.8</td>
<td>48</td>
<td>-</td>
</tr>
<tr>
<td>Beijing [46]</td>
<td>Urban</td>
<td>Jun-Sep.2007</td>
<td>149</td>
<td>73.16</td>
<td>22.80</td>
<td>48.87</td>
<td>47</td>
<td>275</td>
</tr>
<tr>
<td>Tianjin [47]</td>
<td>Urban</td>
<td>Jul-Sep. 2009</td>
<td>87.91</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>130</td>
</tr>
<tr>
<td>Guangzhou [31]</td>
<td>Suburban</td>
<td>Jun 2011-May 2012</td>
<td>40.58</td>
<td>26.06</td>
<td>6.8</td>
<td>11.7</td>
<td>26</td>
<td>-</td>
</tr>
<tr>
<td>Hong Kong [48]</td>
<td>Suburban</td>
<td>Oct-Dec.2007</td>
<td>34</td>
<td>-</td>
<td>-</td>
<td>-</td>
<td>32</td>
<td>-</td>
</tr>
</tbody>
</table>

Table 2. Comparison of VOCs and O$_3$ concentrations at Zibo with observations at other locations in other regions.
higher than in other days, which may have contributed to the very high afternoon O$_3$ concentrations of approximately 160 ppb observed on that day.

To quantify the role of photochemistry formation in the chemical O$_3$ budget, the in situ O$_3$ production rate and destruction rate was calculated by the observation-based model. The in situ day-by-day variations of model-computed O$_3$ production, destruction and net rates for this episode are presented in Fig. 5. The model simulated average O$_3$ production rate (~30.4-41.2 ppb/h) during the 2-5 July 2018 is much higher than that of O$_3$ loss rate (~5.1-8.0 ppb/h), indicative of net O$_3$ production in this episode. Fig. 5 also shows the breakdown of the ozone production and loss pathways simulated in this episode at Zibo. In general, all four case days showed similar O$_3$ formation and destruction mechanisms. The O$_3$ production is dominated by the reaction of NO with HO$_2$, which accounts for ~47% (i.e. 15.2 ppb/h), ~50% (17.7 ppb/h), 55% (23.0 ppb/h) and ~56% (15.3 ppb/h) of the total O$_3$ production for 2-5 July, respectively and the remaining (~53% or 16.6 ppb/h), ~50% (16.0 ppb/h), ~45% (18.6 ppb/h) and ~51% (15.2 ppb/h). O$_3$ production is due to the NO oxidation by RO$_2$, with CH$_3$O alone contributing 10% (3.2 ppb/h), 8% (2.7 ppb/h), 8% (3.3 ppb/h) and 10% (3.0 ppb/h). For the O$_3$ loss terms, the dominant pathway is the NO$_2$+OH reaction to form nitric acid (HNO$_3$) (38% (or 1.4 ppb/h), 26% (1.2 ppb/h), 48% (3.2 ppb/h) and 47% (2.0 ppb/h), respectively), followed by reaction of NO$_2$ with RO$_2$ with average contributions of 12% (or 1.7 ppb/h), 11% (1.6 ppb/h), 9% (2.2 ppb/h) and 11% (1.4 ppb/h), respectively. This indicates the potential intensive formation of organic nitrates and HNO$_3$ at Zibo. Indeed, high levels of fine particulate nitrate in PM$_{2.5}$ in the NCP have been reported by Wen et al. [49], who indicated that controlling NOx and O$_3$ is an efficient way to mitigate nitrate pollution. Wang et al. [50] measured fine particulate nitrated phenols and reflect the important role of secondary formation to nitrated phenols in remote areas of NCP.

In general, the change of observed O$_3$ concentrations is subjected to three factors: in-situ photochemistry (production or destruction), regional transport (import or export) and deposition. Using the observation-based model (OBM), the relative contributions of these factors can be estimated, which has been reported by Xue et al. [26]. Specifically, the change rate of O$_3$ concentrations can be obtained from our observed O$_3$ time series ($R_{\text{meas}}$); the in situ net O$_3$ chemical formation rate ($R_{\text{chem}}$) and deposition rate ($R_{\text{dep}}$) can be estimated by the OBM model; therefore, the regional transport rate ($R_{\text{trans}}$), which includes not only the horizontal and vertical transport but also the atmospheric effects, can be obtained by subtracting (i.e., $R_{\text{trans}} = R_{\text{meas}} - R_{\text{chem}} - R_{\text{dep}}$).

Fig. 6 presents the contribution of in-situ photochemistry, regional transport and deposition at Zibo on 2-5 July, 2018. Similar to Fig. 5, Fig. 6 shows that net O$_3$ production occurred during daytime throughout this episode. In addition, we found an interesting phenomenon that in the early morning, O$_3$ accumulated rapidly owing to the import of O$_3$-rich air aloft from the residual boundary-layer air. The import of O$_3$-rich air aloft contributed ~1.1 to ~4.2 ppbv of the observed O$_3$ accumulation in the morning. Statistically, the daytime average (07:00-18:00 LT) net O$_3$ production rates were calculated as 26.7, 28.6, 33.6 and 24.9 ppb/h for 2-5 July 2018, respectively. Such levels are of the same magnitude as the modeled rates at suburbs downwind of Shanghai and Guangzhou (25 and 30 ppb/h, respectively; [26]) as well as Zhuhai (16 ppb/h, [51]), but are much higher than those obtained at a rural site downwind of Beijing (4.3 ppb/h; [36]) and Mt. Tai (6.4 ppb/h; [25]). This is as expected due to the higher abundance of O$_3$ precursors at Zibo. The locally produced amounts of O$_3$, as estimated by integrating the in situ production rates
Ground-Level Ozone Production...

from early morning to the O\textsubscript{3} maximum, were \~135.2 to \~155.1 ppbv for these four case study days. Such amounts were higher than the observed O\textsubscript{3} build-up of 93.8-144.7 ppbv (see Fig. 6), which were calculated as the increment in O\textsubscript{3} concentrations from early morning to the O\textsubscript{3} maxima. This indicates that photochemical O\textsubscript{3} formation at our study site was sufficiently rapid to lead to the observed O\textsubscript{3} increase during the daytime. Moreover, regional transport made a negative contribution to the O\textsubscript{3} production rate at this site and may had a potential influence to the downwind regions, which is worthy of further study.

**Ozone Formation Regimes**

To diagnose the O\textsubscript{3} formation regimes, we performed model sensitivity simulations during four case study days to determine the O\textsubscript{3}–precursor relationships. O\textsubscript{3}–precursor relationships were determined by calculating the relative incremental reactivity (RIR), which was defined as the percent change in the net ozone production rates per percent change in target precursors (i.e., VOCs, NO\textsubscript{X}, or CO) [52]. In this study, we examine the responses of net ozone production rates to reductions by 20\% of VOCs, NO\textsubscript{X}, CO as well as some particular VOCs groups and species. The model-calculated RIRs for major O\textsubscript{3} precursors and VOC sub-groups at Zibo during the daytime O\textsubscript{3} episodes are summarized in Fig. 7. Overall, the O\textsubscript{3} formation regimes were similar among these four case study days with some minor differences. VOCs have much higher positive RIR values, with daytime averages (07:00-18:00 LT) of 0.77, 0.75, 0.80 and 0.90 on 2-5 July 2018, respectively, suggesting that reducing VOCs may result in effectively reducing O\textsubscript{3} concentrations. The RIRs of NO\textsubscript{X} and CO were positive but small on all episode days, except for 4 July and 5 July, when the RIRs for NOx were negative. This indicates that NOx emissions reductions may even enhance O\textsubscript{3} formation. Therefore, the different RIR values of NOx on these episodes implied that the O\textsubscript{3} formation mechanism varied even at the same site. The ozone formation at Zibo was in a VOC-limited regime or a mixed-control regime. In particular, to further reveal the effects of the VOC subgroups, including alkanes, alkenes and aromatics on O\textsubscript{3} formation, their corresponding RIR values were also calculated and shown (see the right panel of Fig. 7). It can be clearly seen that aromatics exhibited the greatest RIR values within the VOCs, i.e., \~0.45, \~0.43, 0.38 and \~0.49 for the four case study days, respectively, suggesting its dominant role in O\textsubscript{3} production. Hence, the results suggest that the most efficient way to alleviate the O\textsubscript{3} pollution in Zibo is to reduce the aromatics emissions.

Furthermore, the O\textsubscript{3} formation regimes in a diurnal cycle may change with respect to the changes of O\textsubscript{3} precursors [31, 51]. Fig. 8 depicted the diurnal patterns of major O\textsubscript{3} precursors and the VOC sub-groups on episode days at Zibo. This figure clearly illustrates the well-defined diurnal variations of O\textsubscript{3} formation regimes for the polluted industrial environments. It can be seen that most RIRs for O\textsubscript{3} precursors as well as
the VOC sub-groups were positive except for the morning negative values (i.e., 08:00-11:00 LT) of NOx, indicating the switched regime from VOCs-limited to mixed-control regime. Similar phenomena have also been observed in Zhuhai, Guangdong [51], which ascribed the fresh emissions of NOx from nearby vehicular emissions and household activities in the morning.

In particular, with the explicit mechanism of MCM, the OBM model is capable of identifying the O3 precursors at a species level. To further investigate the relative contribution of the individual aromatics species (in total 12 aromatics had been detected in this study) to the formation of O3, a series of sensitivity model simulations that calculate the RIR values of 12 aromatics in the OBM were also performed. Fig. 9 shows the model-calculated 12 aromatics RIRs. Of the 12 aromatics, 1,3,5-trimethylbenzene and toluene were found to dominate the total RIR for aromatics. The summed RIR for these two compounds accounted for approximately 75% to 95% of the total RIR for aromatics, indicating that they exert the greatest influence on in-situ O3 formation at Zibo. In summary, this suggests that reducing anthropogenic emissions of 1,3,5-Trimethylbenzene and toluene would reduce O3 formation in this industrial atmosphere of China.

Furthermore, a positive matrix factorization (PMF) model conducted on the VOC species was used to identify the contributions of different sources to ambient 1,3,5-Trimethylbenzene and toluene. Five resolved source profiles for these aromatics and related VOCs are shown in Fig. S2. The source profiles of

![Fig. 7. The OBM-calculated RIRs for major O3 precursors and the VOC sub-groups on 2-5 July, 2018. The left panel shows the RIRs for NOx, CO and VOCs; the right panel shows the RIRs for the VOC sub-groups.](image1)

![Fig. 8. Average diurnal variation of RIRs for major O3 precursors and the VOC sub-groups at Zibo. The error bars are the standard deviations.](image2)
these aromatics deduced from PMF analysis showed that 1,3,5-Trimethylbenzene and toluene were mainly emitted from solvents used in industrial processes and vehicular exhaust at Zibo, based on a previous study conducted in France [53, 54] and Brazil [19]. Overall, vehicular emissions, industry and solvent usage were the largest sources of these reactive VOCs. Hence, by reducing emissions of aromatics (from industry and vehicles), the photochemical O₃ pollution in the study region would be significantly alleviated.

**Conclusions**

Real-time monitoring of VOCs, O₃ and related trace gases in the industrial region of Zibo from July 1 to 31, 2018 was implemented to improve the understanding of the O₃ pollution, as well as VOCs levels, and contributions of VOCs emissions to O₃. High O₃ concentrations (>93 ppbv) were observed during the early stages of the sampling period. Examinations of diurnal variations illustrate the extensive photochemical O₃ formation. The detailed in situ O₃ photochemical formation and the ozone-precursors relationships during a severe smog episode at Zibo were analysed by an observation-based box model. The reactions of NO + HO₂ and NO + RO₂ dominated the photochemical formation of O₃. The NO₂ destroyed via the reactions with RO₂ and OH, played an important role in the O₃ destruction. Intensive in situ photochemical production during daytime resulted in high O₃ levels, and illustrated the potential to export produced O₃ to the downwind area.

The OBM-derived RIRs result indicated that the O₃ formation was in a VOCs-limited regime or a mixed-control regime in Zibo. The VOCs showed the highest RIR, while NOx showed a slight positive and even negative correlation with O₃ production. This study found that reducing emissions of 1,3,5-trimethylbenzene and toluene would lead to less O₃ production in Zibo, while reducing NOx emissions had little effect and may even increase O₃ pollution. In summary, the results of OBM and PMF analyses indicated that controlling the VOC emissions from industrial, solvent use and vehicle exhaust would likely reduce the photochemical O₃ pollution in the industrial cluster of Zibo and maybe also apply to other polluted industrial regions of China.

**Acknowledgements**

The authors thank Jifeng, Zhao, Chun Ding, Shaoluo, Li and Xuelian, Zhong for their help during the field measurements and data processing. We are grateful to the Master Chemical Mechanism group at the University of Leeds for providing the MCM model. The data analysis and article writing were supported by National Natural Science Foundation of China (No. 42005092), Natural Science Foundation of Shandong Province (No. ZR2020QD058), the Doctoral Research Fund of Shandong Jianzhu University (No. XNBS1936), and the Opening Project of Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention (LAP3) (No. FDLAP19006). These field measurements were funded by Zibo science and technology development plan project (No. 2016kj010009).
Conflict of Interest

The authors declare no conflict of interest.

References

2. INOUE K., TONOKURA K., YAMADA H. Modeling study on the spatial variation of the sensitivity of photochemical ozone concentrations and population exposure to VOC emission reductions in Japan. Air Qual Atmos Heal, 12 (9), 1035, 2019.
Ground-Level Ozone Production... 1897


43. LEUCHNER M., RAPPELLG CL. VOC source–receptor relationships in Houston during TexAQS-II. Atmos Environ, 44 (33), 4056, 2010.


Supplementary Material

Fig. S1. Time series of $O_3$, related pollutants and meteorological conditions at Zibo between 2 and 5 July 2018.

Fig. S2. Source profiles of 1,3,5-Trimethylbenzene, toluene and other VOC species deduced from PMF analysis.
PMF Model

PMF is a factor-based receptor model based on multivariate statistical methods that decompose a matrix of sample data into matrices. In this study, \(744 \times 55\) matrix data (VOCs number \(\times 55\) species) sets were introduced to PMF 5.0 to identify 1,3,5-Trimethylbenzene and toluene sources in Zibo. The detailed model settings and physical parameterizations can be found in the user manual [1]. Five resolved source profiles for 1,3,5-Trimethylbenzene, toluene and related VOCs in Zibo are shown in Fig. S2. The first factor was characterized by high amounts of Alkenes and Aromatics, suggesting its relation to industrial process [2]. Based on the high undecane and n-dodecane levels, which were important markers for burning, factor 2 was identified as burning [3, 4]. The third factor was identified as vehicular emissions, represented by high load of n-butane, iso-pentane, acetylene and toluene. These VOCs had been commonly recognized to be markers of vehicle emissions [4, 5]. As to the fourth factor, the solvent usage was distinguished by high amounts of Ethyl-benzene, 1,3,5-Trimethylbenzene and 1,2,4-Trimethylbenzene [6]. Finally, The high loadings of N-octane combined with 1,2,3-trimethylbenzene indicated that factor 5 was related to LPG/NG usage [4]. Obviously, among these derived profiles of each factor, vehicular emissions were important sources of toluene. Solvent usage as well as industrial process emitted large amount of 1,3,5-Trimethylbenzene. Thus, vehicular emissions, industry and solvent usage were the major sources of 1,3,5-Trimethylbenzene and toluene.

Reference
