

Microorganisms in Soils with High Nickel and Chromium Concentrations in Western Serbia

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Abstract

The occurrence of microorganisms and their relationships with total and available heavy metal concentrations (Cu, Zn, Pb, Cr, Ni, Cd, Hg, and As), as well as main soil chemical characteristics (pH, P, K, Ca, N_{tot}, C_{org}) were evaluated in naturally metalliferous serpentine soils and non-serpentine soils with elevated Ni and Cr concentrations. Significant negative correlations were detected between ammonifiers ($r = -0.545$, $r = -0.371$), oligonitrofiles ($r = -0.478$, $r = -0.458$), total number of microorganisms ($r = -0.363$, $r = -0.393$), and available Ni and Cr. The number of actynomicetes and *Azotobacter* spp. was not affected by heavy metal concentrations, nor any other soil chemical characteristic. Ammonifiers correlated positively with P and K soil content, but negatively with soil C_{org}. High positive correlations were obtained for available metal concentrations and soil C_{org} and N_{tot}, while soil pH only correlated with available Cr and As. Stepwise multiple regression analyses indicated that increased Ni and Cr concentrations had overall little influence on microbial groups, and only ammonifiers were significantly affected by increased available Ni, but P content had an even higher effect on their number in soil.

Keywords: heavy metals, microbial community, soil quality, serpentine, ammonifiers

Introduction

Microorganisms play an intrinsic role in soil formation and maintain fertility, structure, and stability of soils. They are vital to many of the ecological processes that sustain life, such as nutrient cycling, decay of plant matter, consumption and production of trace gases, and

transformation of metals [1]. On the other hand, chemical and physical properties of soils, like quality and amount of soil organic matter, pH, metal concentrations, etc., have a pronounced influence on the dynamics of the microbial community structure and function in soils [2].

Elevated heavy metal concentrations in soil may pose risks and hazards to humans and the ecosystem [3]. Microorganisms are the first group that undergoes any direct or indirect effects of heavy metals in the soil environment [4], and they are good indicators of changes

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in soil biological activity and soil health/quality. Heavy metal toxicity toward soil microorganisms has often been reported, but due to the wide variation in toxicity thresholds (>1000-fold) among studies, the derivation of critical metal concentrations in soils for microbial groups/processes is difficult [5-6]. Soil properties that affect both the toxicity of the metals and soil microbial populations may explain some of these variations. Negative effects of metal contamination on soil respiration, microbial biomass, enzyme activities, and microbial community size and diversity have been documented [7-11]. However, the results indicating no differences between polluted and unpolluted sites in terms of biomass, enzymatic activity, or abundance of some groups of microorganisms have also been published [4, 12-13]. The estimated differences could be caused by limited bioavailability of heavy metals and/or resistance developed by the microbial community to chronic high heavy metal concentrations [13], but also by other physical and chemical interactions in the soil, indicating the complexity of heavy metal influence in the ecosystem. Therefore, the influence of heavy metals and other soil properties on soil microorganisms in field conditions should be analyzed.

Some soils in western Serbia, serpentine soils, are characterized by high levels of heavy metals, mainly Ni and Cr, since they are formed on ultramafic rocks or sediments originating from the ultramafic rocks [14]. In addition, non-serpentine soils with elevated Ni and Cr concentrations are also present in this part of the country. These elevated metal concentrations are probably caused by natural erosion and accumulation of weathering products of ultramafic rocks, but anthropogenic influence cannot be excluded. Our study aimed to evaluate the relationships between the occurrence of some groups of microorganisms and total and available heavy metals, as well as other chemical characteristics of soils in naturally metalliferous serpentine and non-serpentine soils with a wide gradient of elevated Ni and Cr concentrations.

Material and Methods

Site Description

The soils were collected from two regions in western Serbia: Maljen Mountain, which is of ultramafic, serpentine origin (naturally high in Ni, 10 samples); and non-serpentine soils surrounding the town of Čačak (with elevated levels of Ni, 23 samples) (Table 1). On Maljen, Leptosols formed on serpentized harzburgite and Jurassic serpentines dominate, mainly under forests and grass vegetation to a lower extent (Table 2). In the Čačak area Vertisol and Fluvisol soils are dominant, formed on Neogene and alluvial sediments (non-serpentine soils), and arable land and orchards dominate (Table 2). The samples were collected in autumn 2009, from a depth of 25 cm and used for chemical and microbiological analyses.

Soil Microbiological Analysis

The plate count method was used to enumerate the total number of microorganisms, the number of fungi, actinomycetes, and oligonitrophiles in soil samples. The number of microorganisms was determined on the agarised soil extract, fungi on the Chapek medium [15], actinomycetes on the synthetic agar with saccharose according to Krasiljnikoy [16], and oligonitrophiles on the medium according to Fyodorov [16]. The ammonifiers and *Azotobacter* spp. were estimated by the most probable number (MPN) in medium with asparagines as the source of nitrogen for ammonifiers, while liquid medium with mannitol was used for *Azotobacter* spp. [17].

Soil Chemical Analysis

Soil pH was determined with a glass electrode pH meter in 1N KCl (in ratio soil: KCl 1:2.5). Available P and K in soil were determined by the AL-method of Egner-Riehm, while CaCO₃ was determined volumetrically using a Sheibler calcimeter. Soil organic C (C_{org}) and total N (N_{tot}) were determined with an elemental CNS analyser, Vario model EL III (Hanau, Germany). The total content of heavy metals (Cd, Cr, Cu, Hg, As, Ni, Pb, and Zn) was determined by acid digestion (HNO₃) and by plasma emission spectrometry ICAP 6300 ICP (Cambridge, UK). Available forms of Cd, Cr, Cu, As, Ni, Pb, and Zn in soils were extracted by DTPA [18] and determined by ICAP 6300 ICP.

Statistics

Statistical differences on microbial parameters between sites were evaluated using one-way ANOVA. Pearson's correlation and stepwise multiple linear regression were performed to determine the relationship between the number of microorganisms and soil characteristics. The SPSS 16.0 software package was used.

Results

Soil Properties and Heavy Metal Concentrations

Chemical characteristics and heavy metal concentrations of soils in two investigated areas are presented in Table 1. In both investigated areas, in 32 of 33 soil samples investigated, total Ni concentrations were above maximum allowed values for agricultural soils according to national legislation (MAC) (Tables 1-2). Concentrations of Cr above MAC were detected in all serpentine and eight samples from the non-serpentine area. There was a wide gradient of total Ni concentrations in soils from 47 to 1,689 mg kg⁻¹, as well as Cr from 47 to 957 mg kg⁻¹ (Table 2). Although elevated concentrations were detected in both areas, two areas significantly differ in total and available Ni and Cr concentrations. Moreover,

the average total and available Ni concentrations were about 9 and 20 times greater in serpentine compared to non-serpentine soils. The concentrations of all the other heavy metals evaluated (Cu, Zn, Pb, Cd, As (metalloid), and Hg) were below MAC in both areas, with exceptions of single metals in two soil samples (Tables 1-2). Concentrations of total Pb were significantly higher in serpentine soil, while available Cd and total Cu and Hg concentrations, as well as both forms of As were significantly higher in non-serpentine soil. There were no significant differences between soil pH in two areas, which was mostly weak acid reaction. Contents of P and K were significantly higher in non-serpentine, while N_{tot} and C_{org} were significantly higher in serpentine soils. The C/N ratio was favorable in almost all soil samples. The obtained chemical characteristics correspond to typical serpentine soil under the forests [12, 19].

Microbiological Characteristics of Soils

In our research, the number of fungi, actinomycetes, and *Azotobacter* spp. did not significantly differ between the areas, while the total number of microorganisms, ammonifier, and oligonitrofiles were significantly lower in serpentine soil (Table 2). The total microbial number was in the range of $3.67\text{-}38.33 \times 10^6$ and $2.33\text{-}61.33 \times 10^6$ CFU g^{-1} of soil in serpentine and non-serpentine soil, respectively. Although there were no significant differences between *Azotobacter* spp. in two areas, their number in serpentine soil was very low (0-4 MPN g^{-1} of soil) in respect to non-serpentine soil (0-1100 MPN g^{-1} of soil). The number of ammonifiers was in the range of $0.70\text{-}110 \times 10^5$ and $7.5\text{-}140 \times 10^5$ MPN g^{-1} of soil, while oligonitrofiles ranged $10\text{-}82.67 \times 10^5$ and $5\text{-}191 \times 10^5$ CFU g^{-1} of soil for serpentine and non-serpentine soil, respectively.

Table 1. Soil chemical properties and heavy metal concentrations.

	Serpentine soil			Non-serpentine soil				
Chemical properties of soil								
	Min	Max	Mean±SD	Min	Max	Mean±SD	ANOVA	MAC
pH _(KCl)	5.32	5.82	5.61 ±0.18	3.50	7.01	5.80±1.01	ns	
P ₂ O ₅ [mg 100g ⁻¹]	0.10	0.78	0.26±0.23	0.00	58.43	14.25±14.98	**	
K ₂ O [mg 100g ⁻¹]	9.82	21.08	14.93±3.94	13.41	45.00	27.82±11.46	**	
CaCO ₃ [mg kg ⁻¹]	0	0	0	0	8.48	0.852±1.77	ns	
N _{tot} [%]	0.37	0.88	0.47±0.19	0.11	0.44	0.23±0.09	***	
C _{org} [%]	2.17	11.39	6.21±2.36	0.94	4.46	2.38±1.02	***	
Total heavy metal concentrations (mg kg ⁻¹)								
Cu	6	11	10.1±2.33	16	181	36.13±34.13	*	100
Zn	40	119	67.1±21.84	44	123	63.91±16.10	ns	300
Pb	26	94	45±19.81	11	100	30.04±17.99	*	100
Cr	359	957	591.4±159.83	47	164	84.91±35.40	***	100
Ni	770	1689	1166.1±253.87	47	329	132.0±80.03	***	50
Cd	0.2	0.85	0.29±0.20	0.2	0.85	0.32±0.15	ns	3
As	1	20	6.3±5.88	10	33	15.13±5.76	***	25
Hg	0.007	0.031	0.0174±0.0078	0.020	0.282	0.0647±0.0542	*	2
Available heavy metal concentrations (mg kg ⁻¹)								
Cu	0.7	1.9	1.07±0.34	0.8	35.4	4.61±7.17	ns	
Zn	0.6	26.6	5.54±7.48	0.4	9.1	2.36±2.08	ns	
Pb	0.5	18.8	5.37±5.09	0.7	19.9	3.46±3.83	ns	
Cr	0.012	0.048	0.0259±0.0083	0	0.014	0.0059±0.0016	***	
Ni	69	186	118.2±38.19	1.00	29.30	6.60±6.93	***	
Cd	0.04	0.80	0.21±0.21	0.03	0.13	0.10±0.09	*	
As	0.02	0.19	0.07±0.04	0.01	0.47	0.16±0.12	*	

MAC - maximum allowed concentrations for total heavy metals in agricultural soils, serpentine soil (n = 10), non-serpentine soil (n = 23), ns-non significant differences between soils, * P<0.05, ** P<0.01, *** P<0.001

Table 2. Microbiological properties of investigated soils and Ni and Cr total concentrations (mg kg⁻¹).

Land use	Cr	Ni	Total number of microorganisms x10 ⁶ CFU	Fungi x10 ⁴ CFU	Actinomycetes x10 ⁴ CFU	<i>Azotobacter</i> MPN	Ammonifiers x10 ⁵ MPN	Oligonitrofiles x10 ⁵ CFU
Serpentine soil								
Forest	607	1689	12.67	6.33	0.33	0	0.90	49.33
Forest	957	1418	15.33	11.00	16.33	0	3.00	65.67
Forest	694	1260	14.00	10.33	1.67	4	110	44.33
Forest	653	1248	13.00	10.67	2.33	0	1.10	53.00
Forest	359	1091	21.67	20.67	32.67	4	1.10	82.67
Forest	514	1069	16.33	8.00	8.00	0	45	46.00
Meadow	601	1079	16.00	6.67	4.67	0	15	46.00
Forest	524	1045	6.00	13.33	9.67	0	0.90	61.33
Meadow	522	992	3.67	17.33	3.00	4	0.70	10.00
Meadow	485	770	38.33	24.00	7.00	4	9.50	48.33
Mean±SD			15.7±9.46	12.83±6.00	8.56±9.69	1.60±2.06	18.72±34.87	50.66±18.59
Non-serpentine soil								
Orchard	164	329	41.00	14.00	1.67	450	140	46.67
¹² Meadow	177	294	11.33	17.00	6.33	95	45	69.00
Arable	130	248	13.67	9.00	6.00	45	110	92.67
Meadow	110	210	28.00	9.00	7.33	150	140	67.00
Arable	106	195	13.67	7.00	2.00	45	15	95.00
Orchard	121	185	35.67	21.33	2.00	150	20	111.00
Meadow	102	165	61.33	21.33	12.00	450	110	191.00
Arable	73	158	19.33	17.00	5.67	9	25	111.67
Arable	66	146	22.67	6.33	12.00	95	20	118.00
²⁰ Orchard	85	127	19.67	6.33	7.67	0	140	48.00
Arable	84	122	20.00	15.67	5.00	1100	45	67.00
Yard	73	118	31.33	24.00	7.33	25	45	102.00
Meadow	79	92	14.33	20.33	2.33	45	45	89.67
Yard	57	83	25.00	20.00	2.33	95	140	102.67
Orchard	71	79	28.33	15.33	2.33	95	140	102.67
Yard	57	74	26.67	18.00	5.00	95	140	81.67
Meadow	57	72	36.33	25.00	3.33	25	110	83.67
Orchard	56	64	32.67	13.67	0.67	150	140	97.33
Meadow	66	62	25.00	7.67	1.00	25	45	73.33
Arable	60	60	14.33	36.00	2.33	25	7.50	38.33
Orchard	54	59	26.00	23.33	2.33	95	2	89.00
Meadow	58	53	22.33	19.33	10.33	45	140	91.67
Forest	47	47	2.33	9.67	3.00	0	30	5.00
Mean±SD			24.82±12.10	16.21±7.35	4.78±3.89	143.86±240.38	78.80±53.86	85.83±35.33
ANOVA			*	ns	ns	ns	**	**

CFU g⁻¹ soil or MPN g⁻¹ soil; ns-non significant differences between soils, * P<0.05, ** P<0.01, Soil no 12 As = 33 and Pb = 100 mg kg⁻¹, Soil no 20 Cu = 181 mg kg⁻¹

Table 3. Significant Pearson's correlation coefficients between groups of microorganisms and chemical characteristics of soils.

	pH	P ₂ O ₅	K ₂ O	CaCO ₃	N _{tot}	C _{org}	TMic	Fun	Act	Azo	Amm	Olig
P ₂ O ₅	0.499**	1										
K ₂ O		0.548**	1									
CaCO ₃				1								
N _{tot}					1							
C _{org}		-0.375*			0.961**	1						
TMic			0.414*				1					
Fun			0.370*				0.348*	1				
Act									1			
Azo							0.353*			1		
Amm		0.554**	0.377*		-0.356*	-0.424*	0.447*				1	
Olig							0.654**					1
Total metal concentrations												
Cu			0.357*	0.890**								
Zn					0.487**	0.426**						
Pb					0.584**	0.565**						
Cr		-0.409*	-0.487**		0.570**	0.653**	-0.351*				-0.449*	-0.437**
Ni		-0.410*	-0.506**		0.532**	0.654**	-0.369*				-0.474*	-0.427**
Cd					0.389*							
As		0.362*	0.464**									
Hg			0.364*				0.477**					0.643**
Available metal concentrations												
Cu			0.391*	0.895**								
Zn					0.730**	0.717**						
Pb					0.578**	0.532**						
Cr	-0.367*	-0.393*	-0.439*		0.619**	0.666**	-0.393*				-0.371*	-0.458**
Ni		-0.433*	-0.498**		0.734**	0.813**	-0.363*				-0.545**	-0.478**
Cd					0.724**	0.699**						
As	-0.610*		0.416*					0.537*				

TMic- Total microorganisms, Fun- Fungi, Act- Actinomycetes, Azo- *Azotobacter*, Amm- Ammonifiers, Olig- Oligonitrofiles, No of soil samples analyzed n = 33, *- P<0.05, **- P<0.01

Although moderate, significant negative correlations were detected between total As as well as available Ni and Cr concentrations and total microorganisms number, ammonifiers, and oligonitrofiles (Table 3). The numbers of actinomycetes and *Azotobacter* spp. were not affected by heavy metal concentrations, nor any other heavy metal or soil chemical characteristic investigated. Positive correlations between Hg and total number of microorganisms as well as oligonitrofiles were observed, while available As positively correlated with the number of fungi. The concentrations of these metals were far below MAC.

In our study, there was no correlation between pH and any microbiological group investigated. Significant negative correlations were detected only between pH and available Cr and As. Ammonifiers positively correlated to P and K, but negatively correlated with soil C_{org} and N_{tot}, while the number of microorganisms and fungi were in positive correlation with K. Considering other interactions, there were significant correlations between total and available forms of each investigated metal (Table 3). N_{tot} and C_{org} were in significant positive correlation with total and/or available concentrations of Zn, Pb, Cr, Ni, and Cd. Negative correlations were noticed between

Table 4. Stepwise multiple regression of microbial groups (dependent variables) and soil properties (predictors) ($P < 0.05$).

Dependent variable	Predictor	R ²
Ammonifiers	P ₂ O ₅	0.307
	P ₂ O ₅ , Ni ^a	0.421
Total microorganisms	Hg ^t	0.227
Oligonitrofiles	Hg ^t	0.413
	Hg ^t , Pb ^t	0.525
	Hg ^t , Pb ^t , CaCO ₃	0.596
	Hg ^t , Pb ^t , CaCO ₃ , Zn ^t	0.654
Fungi	As ^a	0.288
	As ^a , Cd ^t	0.389

a - available, t - total metal concentrations, R² - square of the coefficient of multiple correlation which indicates the proportion of the variance in the dependent variable that is predictable from the independent variable (predictor), negative influence of predictor is marked in bold

P and K content, and Ni and Cr, while Cu and As correlated positively with K.

Stepwise linear regression analysis is a statistical technique used to investigate and model the relationships between variables from multi-factor data. In our study, stepwise multiple regression analysis was used to explore the relationship between all soil characteristics evaluated and the microbial groups, aiming to identify the factors with the highest influence. The analysis indicated that P and Ni are the most important factors that significantly account for variation in ammonifiers (42% of variation explained), with the prevalence of P (30%) over Ni (12%) (Table 4). The total number of microorganisms was the most significantly affected by Hg (22% of variation explained), while the number of oligonitrofiles was affected with total Hg, Pb, CaCO₃, and Zn (65% of variation explained).

Discussion

Although nickel is one of the essential trace metals for microorganisms, plants, and animals, constant exposure to excessive nickel in soils could strongly affect the inhabitant microorganisms, resulting in adverse effects on ecosystem functioning [20-22]. Chromium is not considered essential for microorganisms and plants, but it is a micronutrient in humans and can be highly toxic when accumulated [23]. The response of soil microorganisms to heavy metals depends on soil properties, heavy metal bioavailability, indigenous soil microbial properties, experimental conditions (laboratory or field study), and period of metal exposure (acute or chronic), etc. In our research, the effects of long-term exposure of soil microbial community to elevated Ni and Cr concentrations (values above MAC) were studied in serpentine

and non-serpentine soils with a wide gradient of total Ni and Cr concentrations. Serpentine (ultramafic) soils due to naturally high concentrations of metals, particularly Ni, Cr, and Co, and low concentrations of Ca, N, P, and K represent a naturally stressed environment [24-25]. Lower microbial density and activity of serpentine areas in respect to non-serpentine outcrops have been reported previously [24]. In our study, despite the significant differences in Ni and Cr and some other heavy metal concentrations between the two areas, only some microbial groups were significantly affected. The number of fungi, actinomycetes, and *Azotobacter* spp. did not significantly differ between the areas, while the total number of microorganisms, ammonifier, and oligonitrofiles were significantly lower in serpentine soil. Generally the number of total microorganisms, oligonitrofiles, and actinomycetes in serpentine soil in our study, corresponded to data by Mandić et al. [26] in soils high in Ni and Cr and similar chemical characteristics. The lower number of bacteria in serpentines of Andaman was detected compared to our study, but the differences could be attributed to a much greater Ni content in soils of Andaman (1,740-8,033.4 mg kg⁻¹ soil) [24]. The total culturable number of different microbial groups of the contaminated soils was decreased, but certain groups of soil microbes were particularly sensitive to long-term contamination, asymbiotic nitrogen-fixers, and heterotrophic bacteria, while actinomycetes and fungi were less affected [27], which corresponds to our results.

Negative effects of an increase of Ni biological availability in contaminated soils on metabolic activities and diversity of microbes have been observed [11, 28-30]. However, microbiological activity is not necessarily inhibited by high available Ni concentrations as indicated in the research of Pessoa-Filho et al. [31], who found the differences in microbial biomass C, and some enzymatic activity between two ultramafic soils with different high Ni concentrations but not with non-ultramafic soil with lower Ni. Some research indicates that the plate counts seem to be a more appropriate technique for detecting the effect of metals on soil bacteria than culture-independent methods [32].

Negative correlation between available Ni and ammonifiers as well as oligonitrofiles obtained in our study was previously reported [26]. Niemeyer et al. [10] reported a negative correlation between total Ni and Cr as well as other heavy metals (Fe, Cu, Mn, Zn, Cd, Pb) and ammonifiers. Previous findings in experimental conditions indicate that asymbiotic N₂ fixers are the most affected microbial group with an increasing concentration of Ni [33]. In addition, adding Ni to soil caused a 60% decrease in the number of oligotrophic bacteria and inhibits nitrification processes by up to 20% in comparison to clean soils [34]. It seems that the microorganisms included in the nitrogen metabolism in the soil (ammonifiers and oligonitrofiles) are among the more sensitive to the heavy metal pollution/concentrations [27, 33].

In agreement with our research, it has been reported that actinomycetes are less sensitive to Ni and other

contamination [4, 26-27]. Long-term nickel pollution favored the prevalence of several dominant bacterial guilds, in particular *Actinobacteria* showed tolerance [21]. The tolerance ability of actinobacteria toward different stressful environmental factors, particularly heavy metals, was recognized and possibilities for its use in bioremediation were discussed [21, 35-36].

In the study, positive correlations of Hg and As and the total number of microorganisms, oligonitrofiles, and fungi were observed, but it should be pointed out that the concentrations of these metals were far below MAC. In some research low concentrations of some heavy metals in the soil may increase enzymatic activity, but they are irreversibly inhibited by high concentrations of these metals [37]. Some heavy metals increased the counts of oligotrophic and copiotrophic bacteria, actinomycetes, fungi, and even ammonifying bacteria, but at the same time decreased the number of *Azotobacter* [38]. The growth of some microbial groups may be stimulated by heavy metals in soils of neutral reaction, but inhibited in acid soils demonstrated [39]. This data indicates that the effect of heavy metals should be interpreted in view of other environmental factors.

Microbial communities are influenced not only by heavy metals, but also by other environmental factors such as soil pH, temperature, moisture, and organic matter quality, etc. [8]. The pH level was often connected with microbiological characteristics and availability of heavy metals in soil [21, 30, 40-42]. In our study, there was no correlation between pH and any microbiological group investigated, while significant negative correlations were detected only between pH and available Cr and As, which is possibly due to the lack of significant variation in pH between samples. Previously, significant correlation between heavy metals and pH was not detected due to faintly acidic soil [11, 43-44]. It is considered that decreasing soil pH is the main factor influencing increased nickel solubility, while clay content, iron- manganese mineral, and soil organic matter are of secondary importance [45].

Considering other interactions, there were significant correlations between total and available forms of each investigated metal (data not shown) [43, 46-47]. N_{tot} and C_{org} were in significant positive correlation with total and/or available concentrations of Zn, Pb, Cr, Ni, and Cd. Deng et al. [11] reported significant correlations between several total heavy metals and soil organic contents, but a negative correlation with microbial activity and metals.

Stepwise multiple regression analysis was widely used to investigate and model the relationship between different soil parameters, but it has recently become more common in evaluation factors affecting soil microbial groups [48-50]. In our study, based on the analysis of all soil parameters investigated only ammonifiers were significantly affected by increased available Ni, but P content had an even greater effect on their number

in soil. Previous studies suggest that changes in soil microbiological functioning in serpentine soils were more related to the soil organic matter than to Ni concentrations [12, 31]. The previous results also indicate that microbial diversity and community structure may be regulated by the availability of critical limiting resources [12, 51]. In our research the P content in soil could be a limiting factor for a number of ammonifiers, since the P content in soils investigated is generally low.

Additionally, the microbial community may adapt to the heavy metal concentrations overcoming the negative effects of contamination [52-54]. None of the microbial parameters (biomass, respiration or respiratory quotient, or diversity) were correlated to the gradient in heavy metals of ultramafic soils, suggesting adaptation by the microbial community to the stress imposed by edaphic conditions [12]. Some studies showed microbial tolerance and adaptation to soil contamination due to the long-term metal exposure and indicated that soil organic matter is a key factor in maintaining high microbial enzyme activities and mitigating metal toxicity [53]. In general, microorganisms from serpentine soils showed a high level of tolerance to trace metals, such as Ni, Co, and Cr. An increase in the Ni concentration of the culture medium led to a significant decrease in bacterial densities of all soils. However, this decrease was less pronounced in the serpentine soils, indicating their adaptation to the toxic levels of Ni that they are normally exposed to [55]. Sometimes it may be difficult to attribute the observed effects on microbial communities to the heavy metal pollution due to their complex interactions in soil ecosystems [8, 30].

Conclusions

We can conclude that despite significant differences in Ni and Cr and some other heavy metal concentrations between serpentine and non-serpentine soils, only some microbial groups were significantly affected. In general, with the exception of ammonifiers, the results of regression analysis indicate the lack of significant effect of total and available Ni and Cr on microbial groups in soil with their high concentrations. This could be connected to previously noted developed tolerance of the microbial community to chronic high heavy metal concentrations. It appears that ammonifiers are the most sensitive microbial group to increased Ni concentrations and they can be perceived as sensitive indicators of its increase in soil ecosystems.

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